

1 Termite mound emissions of CH₄ and CO₂ are primarily determined by
2 seasonal changes in termite biomass and behavior

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18 **Keywords**

19 Carbon dioxide, methane, *Microcerotermes nervosus*, mounds, termites, termite biomass

20

21 **Abstract**

22 Termites are an uncertain component in the global source budgets of CH₄ and CO₂. Large
23 seasonal variations in termite mound fluxes have been reported in tropical savannas
24 which should be accounted for when scaling up to annual budgets. The factors driving
25 these seasonal variations in termite mound fluxes are unknown. This paper aims to
26 explain the processes responsible for these seasonal variations in CH₄ and CO₂ fluxes
27 from the mounds of *Microcerotermes nervosus* (Hill) in Australian tropical savannas.
28 Fluxes of CH₄ and CO₂ measured from termite mound sub-samples in the laboratory were
29 a direct function of termite biomass in those mound sub-samples. Termite biomass in
30 mound sub-samples was 10 fold greater in the wet season as compared to the dry season,
31 and was the main factor responsible for the observed seasonal variations in mound fluxes.
32 When expressed per unit termite biomass, termite fluxes were 1.2 (CH₄) and 1.4 (CO₂)
33 fold greater in the wet season compared to the dry season. However, the slightly greater
34 flux emissions per unit termite biomass in the wet season can only explain a small part of
35 the large seasonal variations in mound fluxes. The role of mound diffusivity, measured
36 indirectly, and seasonal variation in CH₄ oxidation by mound material was negligible in
37 driving the seasonal variations in mound CH₄ fluxes. The short term effect of temperature
38 on flux was significant while that of moisture was not. These results emphasize that
39 seasonal termite population dynamics are likely the main driver for the observed seasonal
40 differences in mound fluxes of CH₄ and CO₂. These findings highlight the need to
41 combine future studies of termite fluxes with detailed studies of termite population
42 dynamics.

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49 1. Introduction

50 Termites are one of the most uncertain components in the global budgets of CH₄ and CO₂
51 (Bignell et al., 1997; Brümmer et al., 2009; Khalil et al., 1990). This uncertainty is
52 mainly associated with scaling up factors, such as global estimates of termite biomass or
53 number of nests (Khalil et al., 1990), and the lack of process-based understanding of CH₄
54 and CO₂ exchange between termites and the atmosphere. An important factor for
55 consideration in the scaling up of fluxes from termites is the large seasonal variations in
56 termite mound fluxes of CH₄ and CO₂. CH₄ fluxes from a mound of *Coptotermes lacteus*
57 in summer were greater than rest of the three seasons combined in the sub-tropical
58 Australia, with large seasonal variations in mound CO₂ fluxes as well (Khalil et al.,
59 1990). In the tropical savannas of northern Australia, CH₄ fluxes from the mounds of four
60 termite species were 5 to 26 fold greater in the wet season as compared to the dry season
61 (Jamali et al., 2010). In an Australian tropical semi-arid woodland, Holt (1987) reported
62 large seasonal variation in CO₂ fluxes from the mounds of *Amitermes laurensis*. These
63 seasonal variations in mound fluxes have mainly been correlated with temperature in the
64 sub-tropics (Khalil et al., 1990), and with moisture in the tropical savannas as there is
65 only a small variation in temperature in the tropical savannas on seasonal scale (Jamali et
66 al., 2010). While the effect of temperature on termite fluxes of CH₄ and CO₂ have been
67 reported (Jamali et al., 2010; Shelton and Appel, 2000; Zimmerman and Greenberg,
68 1983), the effect of moisture is still unknown. However, to our knowledge no process

69 based study has been conducted which could confirm the factors causing these seasonal
70 variations in mound fluxes of CH₄ and CO₂ in the tropical savannas.

71 The observed seasonality in mound fluxes of CH₄ and CO₂ can be caused by a number of
72 different factors, such as emissions per time, termite biomass, termite activity, gas
73 diffusivity of termite mound material and CH₄ uptake by mound material:

74 First, these seasonal variations in mound fluxes could be caused by a seasonal change in
75 *emissions per unit termite biomass*. Environmental factors, such as temperature, moisture,
76 or food quality can change the rates of metabolism and respiration in termites.

77 Second, the seasonal dynamics in mound fluxes of CH₄ and CO₂ could also be caused by
78 *changes in the number of termites per mound*. For example, certain termite species can
79 lose up to 50% of their colony biomass as a result of swarming (Wood and Sands, 1978).

80 Third, seasonal variation in *termite activity* such as foraging outside mounds can result in
81 seasonal variation in mound-based flux measurements. This can have implications on
82 termite flux estimates based on termite mounds alone, as only a fraction of the termites in
83 the colony will be present in the mound whilst the remainder of the termites will be
84 emitting CH₄ and CO₂ elsewhere in the ecosystem.

85 Fourth, termite mound walls are mainly composed of soil and can oxidize a fraction of
86 CH₄ produced by termites inside mounds as a result of methanotrophic activity
87 (Sugimoto et al., 1998). Variable mound water contents across seasons can cause *variable*
88 *CH₄ oxidation rates* (and thus variable mound CH₄ fluxes) as oxidation rates can be
89 influenced by moisture. Seasonal variations in mound fluxes of CO₂ can also be partly
90 because of the effect of moisture and temperature on the respiration of microbial biomass
91 in the mound walls (Holt, 1987).

92 Fifth, seasonal variation in *mound diffusivity* as a result of changing mound water content
93 can also cause seasonal variation in mound fluxes.

94 The main aim of this study was to investigate for the first time the factors causing the
95 seasonal variations in mound fluxes of CH₄ and CO₂ in the tropical savannas. All the
96 experiments were conducted on *Microcerotermes nervosus* which is one of the most
97 common mound-building termite species in northern Australia (Watson and Abbey,
98 1993). The objectives were to investigate the: (1) seasonality in CH₄ and CO₂ emissions
99 per unit termite biomass, (2) seasonality in termite biomass dynamics in mounds; (3)
100 seasonality in mound diffusivity and fluxes of mound material as a result of microbial
101 activity, and (4) short term effect of temperature and moisture on termite (not mounds)
102 fluxes of CH₄ and CO₂.

103

104 2. Methods

105 2.1 Site

106 Field work was conducted in a savanna woodland at the CSIRO Tropical Ecosystems
107 Research Center (TERC 12° 24' S, 130 ° 55' E), near Darwin in northern Australia. The
108 vegetation is dominated by *Eucalyptus tetrodonta* F. Muell and *E. miniata* Cunn. ex
109 Schauer over a ground layer of annual and perennial C4 grasses, and a thick litter layer
110 (Dawes-Gromadzki and Spain, 2003).

111

112 2.2 Field-based flux measurements

113 Five mounds of *Microcerotermes nervosus* (Hill) were repeat-measured for CH₄ and CO₂
114 fluxes between February and December 2009, at intervals of four to six weeks. Mound

115 selection was not random; rather mounds in locations that permitted easy access for
116 measurements were selected. Fluxes were measured using static manual chambers of
117 volume 0.02 m^3 , constructed from polyvinylchloride. A collar was permanently installed
118 around the mounds to a soil depth of 3 cm. A chamber of equal circumference to the
119 collar was carefully placed over the mound and connected to the collar using a ribbon of
120 closed cell foam and several tension spring-clamps. This chamber was then connected on
121 the other end to a Los Gatos Research (LGRTM) Fast Greenhouse Gas Analyzer (FGGA)
122 through a pair of gas tubes and SwagelokTM push-fittings. A LCD screen was attached to
123 the FGGA which displayed the CH_4 and CO_2 concentrations measured at a frequency of
124 1Hz (i.e. one sample per second) for a period of five minutes per chamber. The operation
125 of the FGGA is based on an off-axis integrated cavity output spectroscopy combined with
126 a highly specific narrow band laser for the detection of CH_4 and CO_2 strongly reflective
127 mirrors to obtain a laser path length of $2\text{--}20 \times 10^3 \text{ m}$. Further technical details on FGGA
128 operation can be found in Hendriks et al. (2008). Flux was calculated from the linear
129 change in the concentration of CH_4 and CO_2 in the chamber headspace by multiplying the
130 slope ($\text{ppm}_v \text{ hour}^{-1}$) by the chamber volume (L) and dividing by the mound basal area
131 (m^2). Flux was then corrected for temperature and pressure based on the ideal gas law.

132

133 2.2.1 Auxiliary environmental measurements

134 Mound temperature (T_{mound}) was measured immediately after the mound flux
135 measurement by horizontally inserting a hand held Cole-Palmer[®] stainless steel
136 temperature probe 6 cm into the mound. Mound water content was not directly measured
137 to avoid destruction of the mounds required for repeat measurement of CH_4 and CO_2 flux
138 across the seasons. Instead, soil water content (%) was measured gravimetrically by

139 collecting five soil core samples from the top 6 cm next to each mound using a brass soil
140 sampling ring. These were weighed, oven dried at 105 °C and reweighed. Monthly
141 rainfall (mm) data was obtained from the Darwin Airport meteorological station of the
142 Bureau of Meteorology, Australia; located less than 2 km from the TERC site.

143

144 2.3 Flux and termite biomass measurements in the laboratory

145 Sub-samples (n = 22) from *M. nervosus* mounds were collected in 3 L glass jars and
146 equilibrated at 25°C for five hours in a temperature controlled room at Charles Darwin
147 University, prior to measurement of CH₄ and CO₂ fluxes. Fluxes were measured by
148 connecting the glass jars to the FGGA and observing a linear change in the headspace
149 concentration of CH₄ and CO₂ at a frequency of 1 Hz for a period of 10 minutes. Termite
150 biomass was determined immediately afterwards by breaking down the mound sub-
151 samples and collecting individual termites using forceps. The fresh biomass of workers
152 and soldiers were weighed separately to an accuracy of 10⁻⁴ g. The mean biomass of an
153 individual termite within a caste (i.e. workers, soldiers and alates) was determined by
154 weighing 10 individuals from each caste from most of the mound sub-samples. The
155 volume of mound sub-samples was measured, before breaking and removing the termites,
156 by cling-wrapping the sample in a thin plastic sheet and placing it in a partially water
157 filled calibrated container. The volume of displaced water was subtracted from the
158 chamber volume to calculate the net headspace volume. As hand sorting is a time
159 consuming process, a maximum of only two mound sub-samples were collected and
160 measured each day. This experiment was carried out in both the wet (n = 22) and the dry
161 (n = 22) season and was completed within a two week period for both seasons.

162 Fluxes were also measured from the mound material that was left over after removing the
163 termites from mound sub-samples. These mound material samples were incubated for 20
164 minutes using the same set up as described above for incubating mound-samples
165 containing termites but using 1L glass jars. Seasonal difference in fluxes from mound
166 material would explain the role of CH₄ oxidation and microbial respiration in causing
167 seasonal variations in mound fluxes of CH₄ and CO₂. These fluxes from mound material
168 were subtracted from the gross fluxes of mound sub-samples, measured before removing
169 the termites, for calculating the net CH₄ and CO₂ from termites only.

170

171 2.4 Gas diffusivity measurements of mounds

172 Seasonal difference in gas diffusivity of mound wall was measured indirectly by using
173 the ratio of internal mound CH₄ concentration and mound CH₄ flux. In this experiment,
174 11 mounds of *M. nervosus* were repeat-measured for mound CH₄ flux and internal mound
175 CH₄ concentration in the wet and the dry seasons. Internal mound CH₄ concentration was
176 measured immediately after the mound flux measurements by collecting 20 ml gas
177 samples from mounds using a syringe and tube. These gas samples were immediately
178 transferred to pre-evacuated glass vials (Labco Exetainer) which were analyzed for CH₄
179 concentration (ppm) using an auto-injected gas chromatograph (GC, Shimadzu™,
180 GC17a) at the Creswick laboratories of the University of Melbourne. Seasonal variation
181 in CH₄ flux to internal mound CH₄ concentration ratio, and the consistency of
182 relationship between mound CH₄ flux and internal mound CH₄ concentration across
183 seasons would help explain if there is a seasonal difference in mound wall diffusivity.

184

185 2.5 Effect of temperature and moisture on laboratory termite fluxes

186 The short term effect of temperature on termite fluxes was measured in the laboratory,
187 using mound sub-samples ($n = 5$) of *M. nervosus* that contained termites collected from
188 TERC. These were kept in 3 L glass jars and housed in a temperature controlled room at
189 the Charles Darwin University, NT, Australia. Fluxes were measured at three
190 temperatures, 25°C, 35°C and 15°C, after equilibrating for 6 hours at each temperature.

191 The effect of moisture on termite fluxes was investigated by measuring fluxes before and
192 after placing wet calico cloth pieces in the jars at a constant temperature of 25°C; using
193 the same set up as described for the temperature effect. Fluxes from the wet calico were
194 also measured and subtracted from the total fluxes.

195 2.6 Data analysis and presentation

196 SPSSTM 16.0 was used for the statistical analyses of data. Statistical significance was
197 defined at $p \leq 0.05$, unless otherwise stated. Note that original data was used in all the
198 figures and transformed data was used for statistical tests where necessary as stated.

199 A simple linear regression ($n = 30$) was used for testing the relationship of mean mound
200 fluxes (CH_4 and CO_2), measured in field, with mean mound temperature and mean soil
201 water content.

202 A simple linear regression ($n = 22$) was used for testing the relationship between mound
203 CH_4 flux ($\mu\text{g CH}_4\text{-C m}^{-2} \text{ h}^{-1}$) and internal mound CH_4 concentration (ppm). A paired T-
204 test ($n = 11$) was used for analyzing the significance of difference in the mound CH_4 flux
205 to mound internal CH_4 concentration (ppm) ratio between the wet and the dry seasons.

206 For the fluxes measured from mound sub-samples in the laboratory, a simple linear
207 regression analysis was used for testing the relationship between termite biomass and flux
208 (CH_4 and CO_2) separately for the wet ($n = 22$) and the dry ($n = 22$) seasons. An

209 independent sample T-test (n = 22) was used for analyzing the significance of difference
210 in flux per unit termite biomass between the wet and the dry seasons; data was
211 transformed using ln(flux). An independent sample T-test (n = 22) was also used for
212 testing the significance of difference in termite biomass per unit mound sub-sample mass
213 between the wet and the dry seasons; data was transformed using the log₁₀(termite
214 biomass).

215 For analyzing the effect of temperature on fluxes a Q₁₀ temperature coefficient, which is a
216 measure of the rate of change of a biological or chemical system (in this case CH₄ and
217 CO₂ flux) as a consequence of increasing the temperature by 10°C; was calculated as
218 follows:

$$219 \quad Q_{10} = \left(\frac{F_2}{F_1} \right)^{\frac{10}{(T_2 - T_1)}} \quad (1)$$

220 where $F_{1,2}$ are fluxes at two different temperatures, and T is corresponding room
221 temperature (°C). Q₁₀ was calculated for a temperature range of 15 to 25 °C and 25 to 35
222 °C.

223 Paired sample T-test was used for analyzing the effect of moisture on CH₄ and CO₂
224 fluxes from the mound sub-samples (n = 5); data was transformed using log₁₀(flux).

225

226 3. Results

227 3. 1. Seasonal fluxes measured in field

228 CH₄

229 Mound CH₄ fluxes measured in field from *M. nervosus* were 3.5 fold greater in the wet
230 season compared to the dry season. Mean mound CH₄ flux was 1465 ± 293 μg CH₄-C m⁻²

231 h^{-1} in the wet season and $417 \pm 74 \mu\text{g CH}_4\text{-C m}^{-2} \text{h}^{-1}$ in the dry season (Fig. 1). There was
232 a significant relationship ($R^2 = 0.69$, $p \leq 0.05$) between soil water content and mound
233 CH_4 flux, but no significant relationship between mound temperature and mound CH_4
234 flux (Table 1).

235 CO_2

236 Mean mound CO_2 flux was $601 \pm 98 \text{ mg CO}_2\text{-C m}^{-2} \text{h}^{-1}$ in the wet season and 173 ± 34
237 $\text{mg CO}_2\text{-C m}^{-2} \text{h}^{-1}$ in the dry season, i.e. a 3.5 fold difference (Fig. 1). There was a
238 significant relationship ($R^2 = 0.69$, $p \leq 0.05$) between soil water content and mound CO_2
239 flux, but no significant relationship between mound temperature and mound CO_2 flux
240 (Table 1).

241 The 2009 dry season broke in the Darwin region with a 5.4 mm rainfall event in
242 September (Bureau of Meteorology, 2009). Mound fluxes were measured a few days
243 before and within a few hours after this rainfall event. There was a 10 to 50% increase in
244 mound CH_4 flux and a 10 to 80% increase in mound CO_2 flux as a result of this rain (Fig.
245 2). A paired T-test showed a significant difference ($p \leq 0.05$) in mound fluxes (CH_4 and
246 CO_2) measured before and after the rain event (Fig. 2); data was transformed using
247 $\log_{10}(\text{flux})$. Mean mound temperature and mean gravimetric soil water content was 32.4
248 $^{\circ}\text{C}$ and 5.7% before this rain event and 33.4 $^{\circ}\text{C}$ and 10.0 % after the rain, respectively
249 (data not shown).

250 3.1.1 Measurements for mound diffusivity

251 There was a significant relationship between mound CH_4 flux and internal mound CH_4
252 concentration ($R^2 = 0.85$; $p \leq 0.01$) regardless of season (Fig. 3a). The difference in
253 'mound CH_4 flux to internal mound CH_4 concentration ratio' between the wet and the dry
254 season was not significant (Fig. 3b).

255

256 3.2 Fluxes measured in laboratory from mound sub-samples

257 CH₄

258 There was a strong and significant positive linear relationship between termite biomass
259 and CH₄ flux both in the wet ($R^2 = 0.81$, $p \leq 0.001$) and the dry ($R^2 = 0.86$, $p \leq 0.001$)
260 season (Fig. 4a). Mean CH₄ flux was $9.9 \pm 0.8 \mu\text{g CH}_4\text{-C g termite}^{-1} \text{ d}^{-1}$ in the wet season
261 which was significantly greater ($p \leq 0.01$) than the $8.1 \pm 0.6 \mu\text{g CH}_4\text{-C g termite}^{-1} \text{ d}^{-1}$ in
262 the dry season (Table 2). Thus, mean CH₄ flux expressed per unit biomass was 1.2 fold
263 greater in the wet season than the dry season. CH₄ fluxes from the mound material after
264 the termites had been removed were negligible both in the wet and the dry season (data
265 not shown). We did not observe a linear change in CH₄ concentration in jars during
266 incubation of mound material which indicates very low methanotrophic or methanogenic
267 activity in the mound material regardless of season.

268

269 CO₂

270 There was a significant positive linear relationship between termite biomass and CO₂ flux
271 both in the wet ($R^2 = 0.85$, $p \leq 0.001$) and the dry ($R^2 = 0.91$, $p \leq 0.001$) season (Fig. 4b).
272 Mean CO₂ flux was $3.7 \pm 0.8 \text{ mg CO}_2\text{-C g-termite}^{-1} \text{ d}^{-1}$ in the wet season which was
273 significantly greater ($p \leq 0.01$) than the $2.7 \pm 0.2 \text{ mg CO}_2\text{-C g-termite}^{-1} \text{ d}^{-1}$ in the dry
274 season (Table 2), i.e. 1.4 fold greater on a per unit biomass basis. CO₂ fluxes from the
275 mound material (microbial respiration), after removing the termites, were greater in the
276 wet season as compared to the dry season (data not shown).

277

278 3.3 Seasonal variation in termite biomass in mounds

279 As determined from mound sub-samples, there was a significant relationship between
280 termite biomass and mound mass ($R^2 = 0.46$, $p \leq 0.001$) in the wet season but not
281 significant in the dry season (Fig. 5). Termite biomass was 35.0 ± 3.8 g-termite kg-
282 mound⁻¹ in the wet season and 3.6 ± 0.9 g-termite kg-mound⁻¹ in the dry season (Table 2).
283 Thus, mean termite biomass in mound sub-samples was 10 fold greater in the wet season
284 as compared to the dry season. In the wet season, smaller mound sub-samples were
285 collected than in the dry season because of the greater termite biomass density and
286 therefore the time required for separation and removal.

287

288 Soldiers comprised only 5 to 6% of the total termite biomass in a mound, with workers
289 and alates comprising the rest (Table 2). The proportional contribution of workers and
290 alates was not determined because of their similar physical appearance. Mean mass of an
291 individual worker was similar in the wet (1.34 ± 0.04 mg) and the dry (1.41 ± 0.07 mg)
292 seasons, as was the mean mass of an individual soldier in the wet (1.87 ± 0.02 mg) and
293 the dry (1.91 ± 0.11 mg) seasons (Table 2). Mean mass of an alate could only be
294 measured in the dry season (2.8 ± 0.05 mg) as winged alates leave the mounds early in
295 the wet season (Table 2). Thus mass per termite was in the order of alate > soldier >
296 worker.

297

298 3.4 Effect of temperature and moisture on flux

299 For CH₄, the Q₁₀ was 4.6 between 15 and 25 °C and 1.2 between 25 and 35 °C (Fig. 6).

300 For CO₂, the Q₁₀ was 5.4 between 15 and 25 °C and 1.4 between 25 and 35 °C (Fig. 6a).

301 The difference in CH₄ and CO₂ fluxes measured before and after adding moisture to the
302 jars was not significant (Fig. 6b), although there was an increase in termite activity and
303 gallery construction.

304

305 4. Discussion

306 4.1 Seasonal dynamics in termite mound biomass

307 This study demonstrates for the first time that seasonal variations in fluxes of CH₄ and
308 CO₂ from termite mounds in tropical savannas are mainly caused by the seasonal
309 variation in termite biomass in those mounds. We found a 10 fold increase in termite
310 biomass in mound sub-samples in the wet season as compared to the dry season (Table
311 2). We suggest that this was the main factor causing the seasonal variations in mound
312 fluxes measured in the field which were 3.5 fold in this study (Fig. 1) and 8-9 fold in
313 Jamali et al. (2010), as fluxes of CH₄ and CO₂ from mound sub-samples were a function
314 of termite biomass in those mound sub-samples (Fig. 4).

315 There are three probable explanations for the observed seasonal dynamics of *M. nervosus*
316 termite biomass in these mounds:

317 **i) Seasonal dynamics in termite mound population as part of the reproductive cycle**

318 The literature on the life cycle of Australian termites is scarce. For *M. nervosus*, like most
319 tropical species of family Termitidae, swarming usually occurs with the onset of rains
320 between October and December (Hill, 1942); during which time winged alates establish
321 new colonies (Nutting, 1969). Termite colony biomass can be reduced by up to 50% as a
322 result of swarming (Wood and Sands, 1978), but this varies for different termite species
323 (Lepage and Darlington, 2000; Nutting, 1969). Generally, swarming is immediately
324 followed by egg production, which peaks during the wet season (Matsuura et al., 2007).

325 Mature termite populations peak in the dry season, followed by swarming with the onset
326 of rains (Noirot, 1969). However, this suggested lifecycle pattern does not concur with
327 our observations that the greatest CH₄ and CO₂ fluxes and greatest termite biomass in *M.*
328 *nervosus* mounds occur in the wet season (Jamali et al., 2010).

329

330 **ii) Seasonal pattern in termite foraging activity**

331 One possible explanation may be that there are a large number of termites foraging
332 outside the mound in the dry season compared to the wet season, which could explain
333 apparent lower termite numbers in the mound. This is supported by suggestion that
334 termite foraging activity is governed by the energy and protein needs of the colony
335 (Buxton, 1981) which, in tropical areas is greatest in the dry season during nymphal
336 (alate) maturation (Lepage and Darlington, 2000). For the *Macrotermes* species (Bodot,
337 1967; Lepage, 1982; Wood et al., 1977) and *Trinervitermes geminatus* (Ohiago, 1979)
338 studied in African savannas, the peak of foraging activity always occurred in the dry
339 season. This has been suggested as the main reason for smaller termite population in the
340 dry season in the mounds of *Trinervitermes ebenerianus* in the Nigerian savannas (Sands,
341 1965). In a humid tropical forest of Cameroon, Dibog et al. (1998) observed greater
342 termite abundance and species richness in the soil (not mounds) in the dry season as
343 compared to the wet season. Furthermore, the abundance and species richness of these
344 termites was significantly and negatively correlated to the amount of rainfall 48 hours
345 preceding termite sampling. In our study, we observed an increase in CH₄ and CO₂ flux
346 from termite mounds following the ‘break of rains’ after the 2009 dry season (Fig. 3).
347 This response of mound flux to rainfall may be associated with termites being restricted

348 to the mounds due to the wet conditions outside mounds not being suitable for foraging
349 (Dawes-Gromadzki and Spain, 2003).

350 **iii) Seasonal vertical movement of termites**

351 Another theoretical explanation could be vertical movements of termites within the
352 mound. Water requirements for termites are generally very high as most species are
353 poorly protected against dehydration (Collins, 1969). Termites usually maintain high
354 humidity levels within the mound (Noirot, 1970). However, in the dry season when the
355 desired humidity level cannot be sustained, the upper part of the mound is often left
356 empty (Noirot, 1970). Thus, in the dry season the majority of some termite populations
357 resides in the lower and/or underground sections of the mound, where conditions are
358 more moist and humid (Noirot, 1970; Noirot and Darlington, 2000). If this was the case
359 for *M. nervosus* it could have resulted in the smaller termite biomass observed in our
360 mound sub-samples during the dry season compared to the wet season as we have only
361 sampled the aboveground portion of mounds. However, we know of no evidence that
362 there are significant diffusive barriers within the mound which could affect CH₄ and CO₂
363 fluxes because of termite presence in a particular section of mound.

364 Further experimentation involving sampling from both aboveground and belowground
365 mound portions will confirm whether the observed seasonal change in termite population
366 is real or only apparent because of termite activity and termite movement within the
367 mound and to subterranean chambers. The latter case will have significant implications
368 on the estimates based on mound fluxes alone as termites will be emitting CH₄ and CO₂
369 elsewhere in the ecosystem.

370

371 4.2 Seasonality in fluxes per unit termite biomass

372 Seasonal variation in flux per unit termite biomass played only a small role in causing the
373 seasonal variations in mound fluxes of CH₄ and CO₂. The magnitude of the seasonal
374 variation in flux per unit termite biomass was much smaller than the magnitude of
375 observed seasonal variation in mound fluxes measured in the field which was 3.5 fold
376 (CH₄ and CO₂) in this study and 8-9 fold (CH₄ only) in a recent study (Jamali et al.,
377 2010). This seasonal variation in flux per unit termite biomass may be attributed to insect
378 adaptation to xeric conditions, as metabolism and respiration processes can be an
379 important source of water loss (Bartholomew et al., 1985; Edney, 1977; Lighton, 1990).

380

381 4.3 Seasonality in mound diffusivity and fluxes from mound material

382 Mound CH₄ fluxes measured in the field were strongly related to CH₄ concentration
383 inside mounds (Fig. 5a) regardless of season. The difference in the ratio of ‘mound CH₄
384 flux to internal mound CH₄ concentration’ between wet and dry seasons was not
385 significant (Fig. 5b), thus ruling out the role of changing mound wall diffusivity as a
386 driving mechanism in the seasonal variations of mound fluxes. If at all, mound diffusivity
387 is likely to be reduced in the wet season as a result of surface moisture restricting pore
388 continuity in the outer mound wall and therefore cannot explain the greater mound fluxes
389 in the wet season.

390 CH₄ fluxes from the termite mound material were negligible both in the wet and the dry
391 season. We cannot rule out the possibility of CH₄ oxidation by mound material which can
392 be better quantified using long term incubations and isotopic techniques (Sugimoto et al.,
393 1998). However, the absence of measurable CH₄ oxidation in mound material, also
394 reported elsewhere (Bignell et al., 1997), means that this process is unlikely to cause a
395 significant variation in seasonal mound CH₄ fluxes. CO₂ fluxes as a result of microbial

396 respiration from mound material partly contributed towards causing seasonal variations in
397 mound fluxes of CO₂; however, we did not directly quantify their exact contribution.

398

399 4.2 Effect of temperature and moisture on flux

400 There was a positive correlation between temperature and termite fluxes of CH₄ and CO₂.
401 However, temperature fluctuations in tropical savannas are mainly observed on a diurnal
402 basis (day/night) whereas the seasonal differences (wet/dry season) of mean temperatures
403 are rather small. Hence, temperature would not be a major driver for the observed
404 seasonal changes in CH₄ and CO₂ fluxes.

405 The short-term effect of moisture on termite fluxes was not significant despite greater
406 termite activity after the addition of a source of moisture (Fig. 6b). This further supports
407 the argument that seasonal variations in mound fluxes are principally driven by seasonal
408 dynamics in termite population rather than the change in flux per unit termite biomass. It
409 also suggests that an immediate response of mound fluxes to rainfall (Fig. 4) is because
410 of termites being restricted to mounds and not because of any effect on their gut biology,
411 metabolism or physiology.

412

413 5. Conclusions

414 Large seasonal variations in mound fluxes of CH₄ and CO₂ are mainly caused by the
415 seasonal dynamics in termite biomass in mounds. Termites emit slightly greater CH₄ and
416 CO₂ per unit termite biomass in the wet season as compared to the dry season but this
417 does not account for the large seasonal differences observed for mound fluxes of CH₄ and
418 CO₂. Mound diffusivity and CH₄ uptake by methanotrophic bacteria in the mound

419 material play a negligible role and do not influence seasonal variations in mound fluxes.
420 These results emphasize that termite population dynamics are the main driver for the
421 observed seasonal differences in CH₄ and CO₂ fluxes from termite mounds. Although our
422 results could not confirm which processes determined termite population size inside the
423 mound they highlight the need to integrate future studies of termite fluxes with detailed
424 studies of termite population dynamics.

425

426 Acknowledgments

427 This research was supported by the Australian Research Council, Linkage Grant
428 LP0774812. Jamali was supported by an AusAID postgraduate scholarship. We are
429 thankful to Gus Wanganeen from CSIRO Ecosystem Sciences, Darwin for identifying the
430 termite species. We are also thankful to Claire Petit and other students at CSIRO
431 Ecosystem Sciences, Darwin for their help in termite sorting.

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532 Tables

533 Table 1: Relationship of mound fluxes measured in the field with mound temperature and
 534 soil water content as determined by a simple linear relationship

Variable	CH ₄		CO ₂	
	R ²	p	R ²	p
Mound temperature	0.07	n.s	0.10	n.s
Soil water content	0.69	≤ 0.05	0.69	≤ 0.05

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537 Table 2: Seasonal dynamics in flux (per unit termite biomass) and termite biomass in
 538 mound sub-samples of *M. nervosus* as measured in the laboratory

	Wet season	Dry season	Difference (p)
Flux			
CH ₄ (μg CH ₄ -C g-termite ⁻¹ d ⁻¹)	9.9 ± 0.8	8.1 ± 0.6	≤0.01
CO ₂ (mg CO ₂ -C g-termite ⁻¹ d ⁻¹)	3.7 ± 0.8	2.7 ± 0.2	≤0.01
Biomass			
Mean biomass (g-termite kg-mound ⁻¹)	35.0 ± 3.8	3.6 ± 0.9	≤0.01
Mean mass of a worker (mg)	1.34 ± 0.04	1.41 ± 0.07	
Mean mass of a soldier (mg)	1.87 ± 0.02	1.91 ± 0.11	
Mean mass of an alate (mg)	-	2.8 ± 0.05	
Soldiers (% total biomass)	6	5	
Non-soldiers (% total biomass)	94	95	

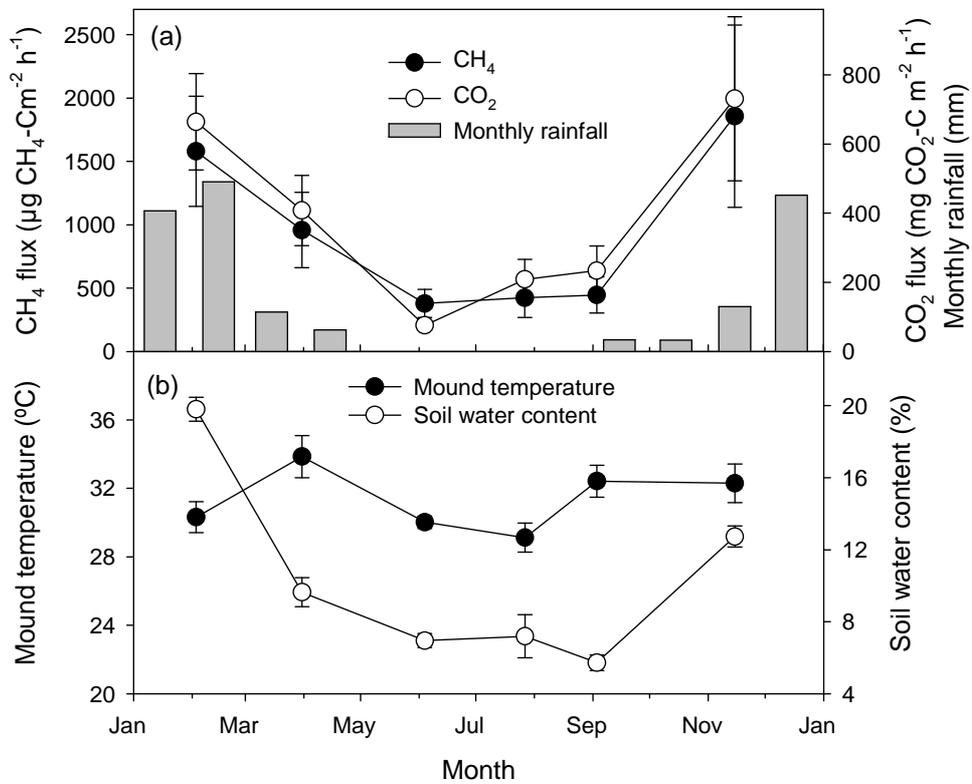
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546 Figure 1: a) Seasonal CH₄ and CO₂ fluxes from five termite mounds of *Microcerotermes*547 *nervosus* measured at TERC site and monthly rainfall from the meteorological station at

548 Darwin Airport, and b) seasonal mound temperature and gravimetric soil water content

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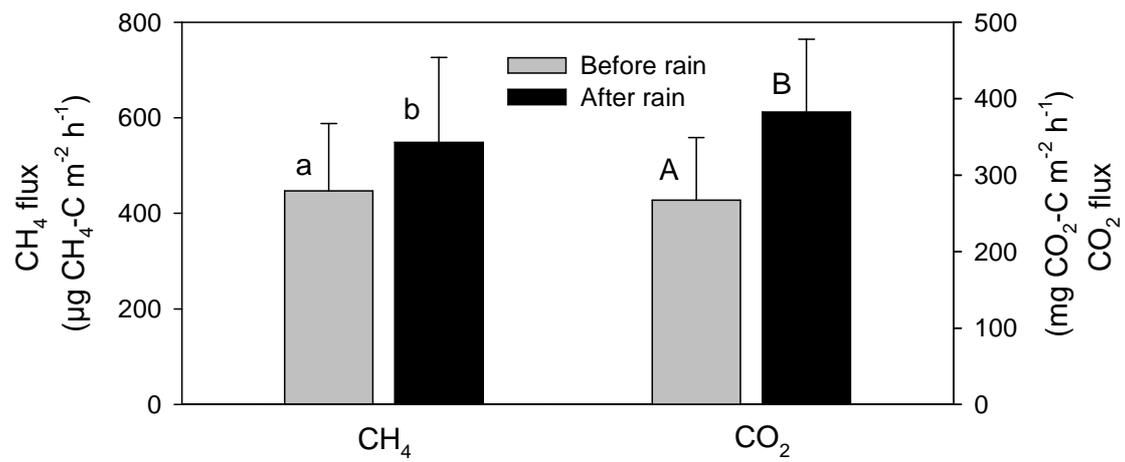
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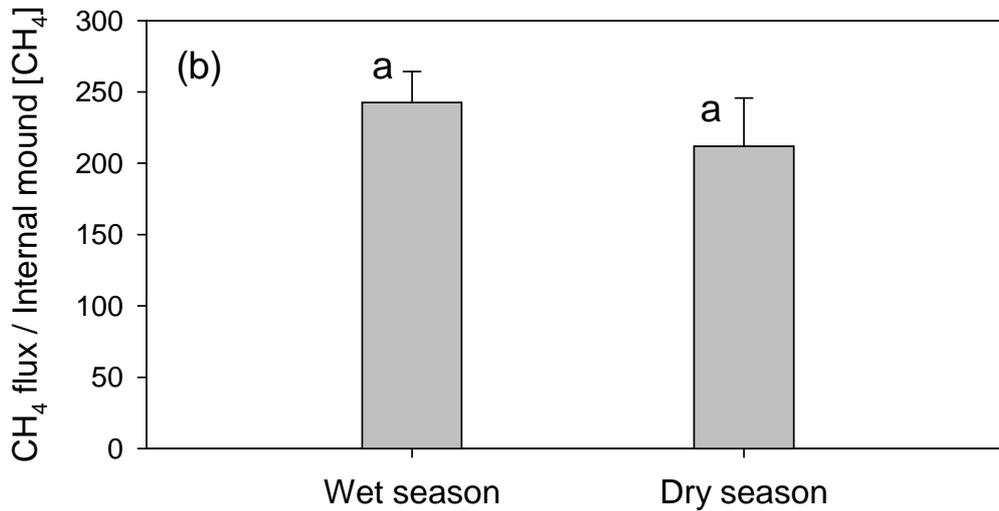
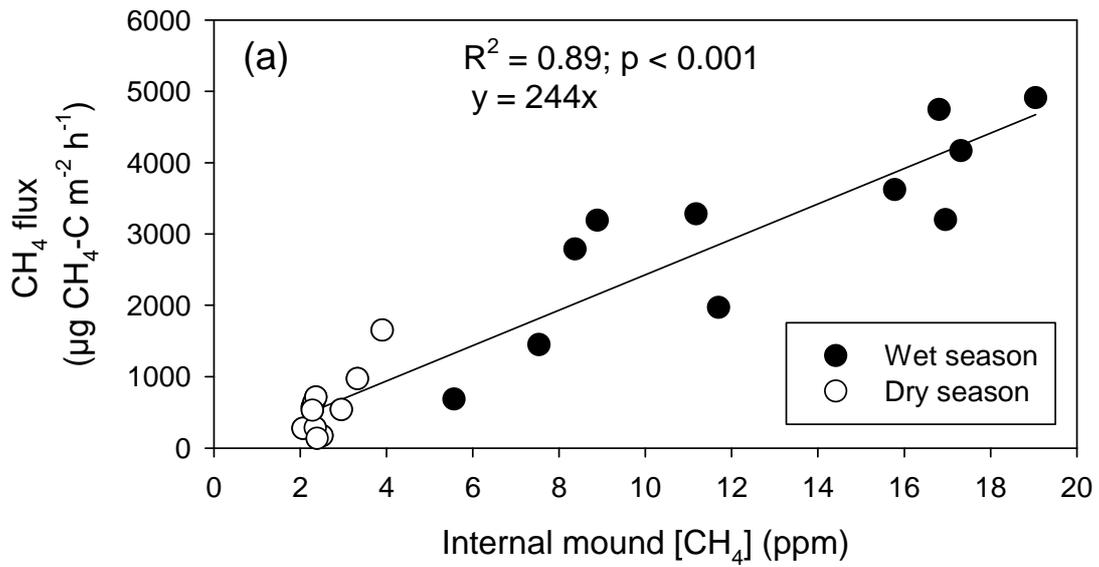
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558 Figure 2: Mean mound fluxes of CH₄ and CO₂ (n = 5) measured from mounds of *M.*
 559 *nervosus* before and after the ‘break of rains’ (5.4 mm) in late dry season 2009 measured
 560 at TERC; error bars are standard error of the mean; case-wise letters show the
 561 significance of difference (p ≤ 0.05)



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563 Figure 3: (a) Simple linear regression between mound CH_4 flux and internal mound CH_4

564 concentration (ppm); (b) seasonal variation in 'mound CH_4 flux ($\mu\text{g CH}_4\text{-C m}^{-2} \text{h}^{-1}$) to

565 internal mound CH_4 concentration (ppm) ratio' repeat-measured in the wet and the dry

566 seasons; letters on top of bars show the significance ($p \leq 0.05$) of difference between the

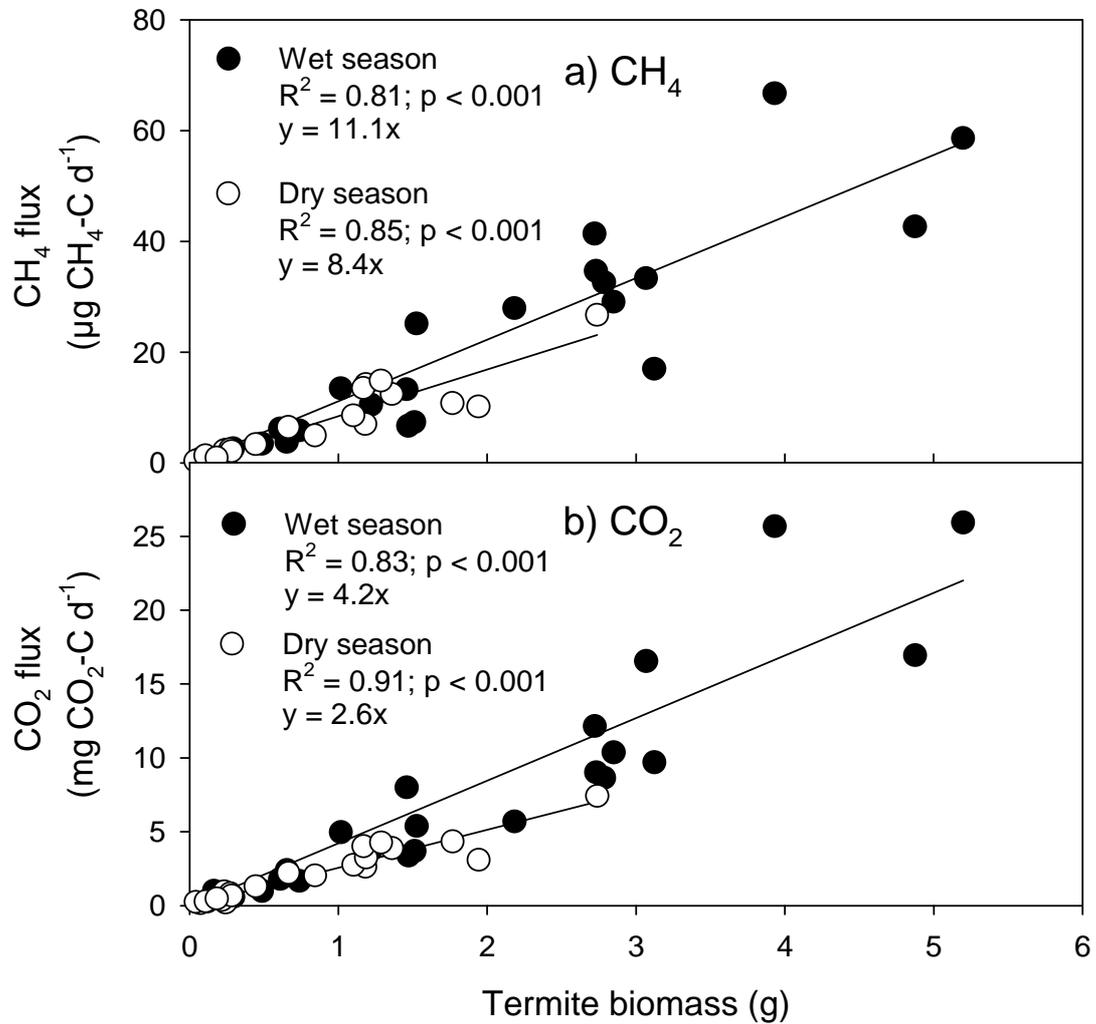
567 wet and the dry seasons

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574 Figure 4: Simple linear regression analyses of: a) fresh termite biomass of *M. nervosus*

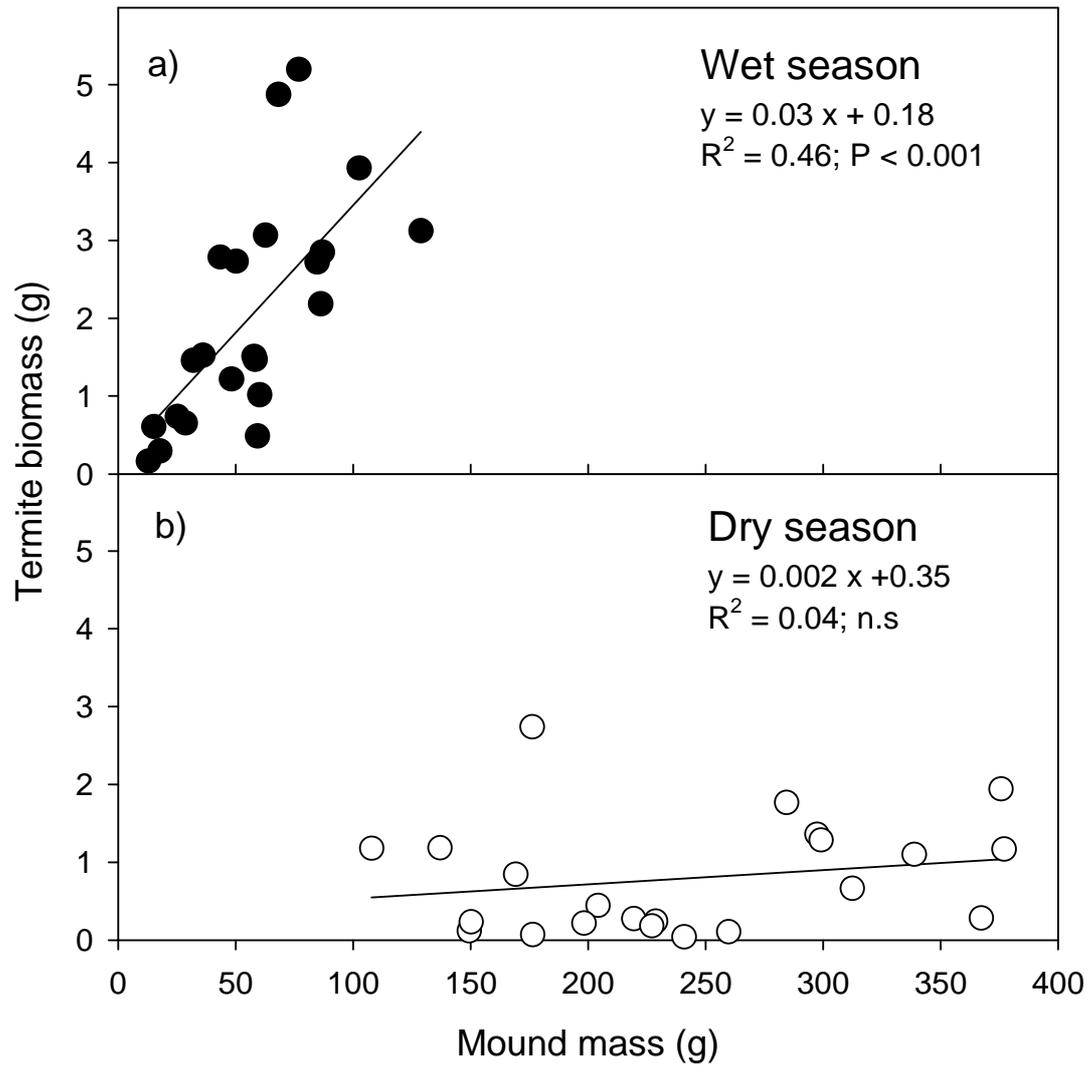
575 and CH₄ flux, and b) fresh termite biomass of *M. nervosus* and CO₂ flux in the wet and

576 the dry seasons of 2009 as measured from fresh mound sub-samples in the laboratory.

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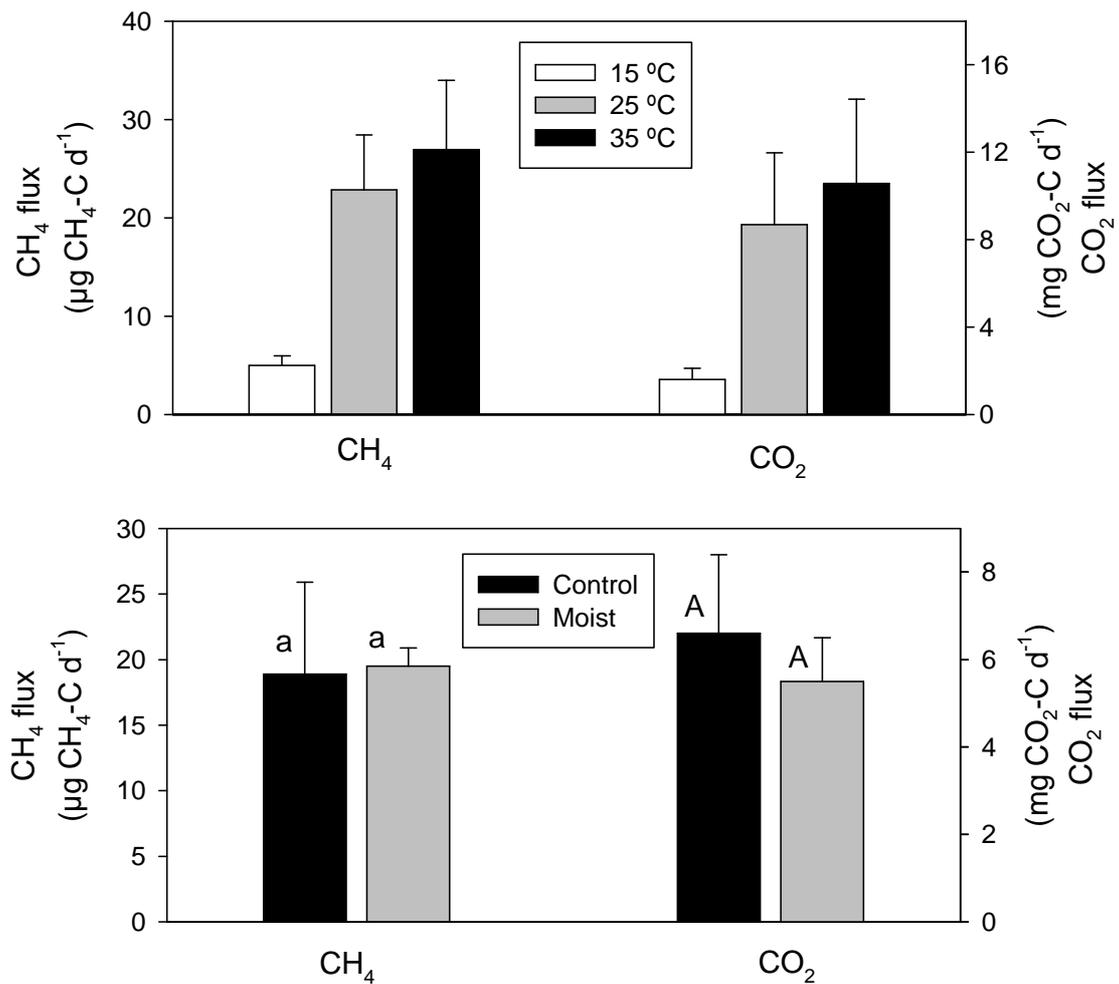
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582 Figure 5: Simple linear regression analysis of mound mass (sub-samples) and termite

583 biomass in the: a) wet season, and b) dry season of 2009, for *M. nervosus*.

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586 Figure 6: (a) Mean CH₄ and CO₂ fluxes measured from mound sub-samples (n = 5)
 587 containing termites incubated at 15 °C, 25 °C and 35 °C; error bars are standard errors of
 588 the mean; case-wise letters on top of the bars show the significance of differences in
 589 fluxes measured at three different temperatures; for CH₄, Q₁₀ values were 4.6 and 1.2
 590 between 15 and 25 °C and between 25 and 35 °C, respectively; for CO₂, Q₁₀ values were
 591 5.4 and 1.4 between 15 and 25 °C and between 25 and 35 °C, respectively.

592 (b) Mean fluxes of CH₄ and CO₂ measured at 25 °C from five mound sub-samples
 593 containing termites; before and after adding moist calico material pieces; case-wise letters
 594 on top of the bars show the significance of variations.

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