Analysis of Climate Change Impacts on the Deterioration of Concrete Infrastructure

*Part 3: Case Studies of Concrete Deterioration and Adaptation*
This report is confidential before public release
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Part 2: Modelling and Simulation of Deterioration and Adaptation Options;
Part 3: Case Studies of Concrete Deterioration and Adaptation.

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1. **INTRODUCTION**

The residual service life of existing concrete structures is largely determined by its deterioration over time. The deterioration rate of concrete structures depends not only on the construction processes employed and the composition of the materials used in the construction process, but also on the current as well as past environment. Meanwhile, climate change may alter this environment in the future, especially in the long term, causing more acceleration of deterioration processes and consequently affecting the safety and serviceability of existing concrete infrastructure. In particular, many existing concrete structures, for which the design has not taken into account the effect of changing climate, are likely to suffer from more decreased durability as a consequence of climate change and incur more damage and maintenance cost. Considering the amount of existing concrete infrastructure, the potential impacts of the climate change cannot be ignored.

As indicated in the previous part report, the deterioration of concrete can be affected directly or indirectly by climate change impacts, in association with the change in carbon dioxide (CO₂) concentration, temperature and relative humidity, as shown in Table 1-1. The climate-related deterioration of concrete structures is mostly caused by the infiltration of deleterious substances from the environment, for example carbon dioxide and chloride, which causes reinforcement corrosion.

Understanding the implications of climate change on existing concrete structure is vital for effective decision-making in asset management to protect concrete buildings and infrastructure that underpin human settlements and the economy. In practice, the
durability and serviceability of concrete structures are maintained via routine inspection, maintenance and replacement. Climate change impacts needs to be considered in both the application and effectiveness of maintenance and replacement regimes. Considering the extensive uncertainties and the limited knowledge of future climate, simulation is once again deemed as an effective approach that may provide insights into how likely and how much the future climate would impact on existing concrete structures. More importantly, taking the precautionary principle, simulations may inform the necessary extent of change in maintenance required to maintain the safety, serviceability and durability at given the likely climate change scenarios.

Table 1-1. Factors and consequences of climate change in association with concrete structures

<table>
<thead>
<tr>
<th>Climate Change</th>
<th>Implications</th>
</tr>
</thead>
<tbody>
<tr>
<td>Increase of carbon concentration</td>
<td>Elevated carbon concentration accelerates carbonation and increases carbonation depth in concrete: this increases the likelihood of concrete structures exposed to carbonation induced reinforcement corrosion initiation and structural damage</td>
</tr>
<tr>
<td>Change of temperature</td>
<td>Elevated temperature accelerates carbonation, chloride penetration and corrosion rate of reinforcement that exacerbates the corrosion initiation and structural damage</td>
</tr>
<tr>
<td>Change of humidity</td>
<td>Lowered relative humidity may reduce or even stop carbonation and chloride penetration in the area with yearly average RH currently just above 40-50%, while increased humidity may result in them occurring in the regions where they are now negligible.</td>
</tr>
</tbody>
</table>

This report will focus on the modelling and simulations for existing concrete infrastructure, and provide an effective way to explore adaptation options to reduce or mitigate the climate change impacts by case studies of concrete bridges and port structures.

In Chapter 2, we will establish methodology and present modelling and simulation for existing concrete deterioration based on field testing data integrated with probabilistic approaches. Models for the simulation of various adaptation measures for chloride and carbonation induced corrosion will also be introduced in the chapter. Using modelling and simulation, cases studies will be carried out in Chapter 3 in relation to the climate change impact on existing concrete bridges constructed in the period of pre-1959, 1959-1970, 1971-1994 in Sydney, Southern, Northern as well as Hunter regions in NSW that are located within a temperate climate zone in Australia. The impact on chloride-induced corrosion will be represented by changes in mean corrosion initiation probability, corrosion damage probability and mean rebar loss. Climate change impact on carbonation-induced corrosion will also be addressed through the same concrete
structures in NSW. In Chapter 4, a similar assessment will be carried out for concrete slabs and columns of port structures in the Port of Townsville that is located within a tropical climate zone in Australia. In Chapter 5, adaptation options to counteract or mitigate climate change impact on both chloride and carbonation induced corrosion will be discussed, and their cost and benefit will be presented in relation to the mitigation of climate change impact, increase of adaptive capacity and offset of adaptation effectiveness loss. Meanwhile, adaptation effectiveness diagrams will be developed for each case for decision-making on the selection of adaptation options in order to most cost-effectively maintain the durability and serviceability of bridges and port structures, which can be further extended to assess other concrete infrastructure.
2. SIMULATION AND MODELLING OF EXISTING CONCRETE STRUCTURES

Simulation and modelling of carbonation and chloride induced corrosion have been discussed in great detail in the second part of the report. These simulations and models are based on the assumption that the design of concrete structures meets the Australian standard of AS3600 – Concrete Structures. Although these simulations and models are able to effectively project the overall performance of general deteriorating concrete structures from the aspect of a new design with consideration of changing climate, they could be inaccurate for the assessment of the performance of an individual existing concrete structure that can be different in practice as a result of uncertainties in construction and local environment as well.

The simulation of carbonation and chloride penetration induced corrosion of existing concrete is implemented by a conventional probabilistic approach on the basis of Monte-Carlo simulation, as demonstrated in the flow chart described in Figure 2-1. This approach considers environmental variables and their uncertainties, including concentration of carbon dioxide, yearly mean temperature and relative humidity. Different from the simulation described in the part 2 report, modelling of deterioration of existing individual concrete structures under changing climate should be calibrated by field testing data, as illustrated in the flow chart, so that corresponding climate adaptations can be properly developed. In this chapter, the procedure to modify the models developed in the previous report will be introduced to consider the characteristics of specific concrete structures. Meanwhile, the modelling of adaptations will also be developed for the assessment of their effectiveness to reduce the impact of climate change.
2.1 Carbonation-Induced Corrosion Modelling with Calibration by Field Testing Data

As discussed in the Part 2 – Modelling and Simulation of Deterioration and Adaptations, carbonation depth at calendar year $t$ is described by

$$x_c(t) \approx \sqrt{\frac{2f}{C(t)}} \frac{D_{CO_2}(t)}{a} \int_{2000}^{t} C_{CO_2}(t) dt \left( \frac{t}{t-1999} \right)^{n_a} \quad t \geq 2000 \quad (2-1)$$

$$D_{CO_2}(t) = D_1(t-1999)^{-n_d} \quad a = 0.75 C_e \alpha_{H} \frac{M_{CO_2}}{M_{CaO}} \quad \alpha_{H} \approx 1 - e^{-3.38w/c} \quad (2-2)$$

where $t_o$ is the reference period (e.g., 1 year), $t$ is time in years, $C_{CO_2}(t)$ is the time-dependent mass concentration of ambient $CO_2$ (10$^{-3}$ kg/m$^3$) using the conversion factor 1 ppm (or ppmv) = 1.9×10$^{-6}$ kg/m$^3$; $D_{CO_2}(t)$ is $CO_2$ diffusion coefficient in concrete; $D_1$ is
CO₂ diffusion coefficient after one year, are based on $T=20^{\circ}$C and RH=65%; $n_d$ is the age factor for the CO₂ diffusion coefficient as shown in Table 2-1, and used for estimate other intermediate values by linear interpolation; $C_c$ is cement content (kg/m³); $C_{CaO}$ is CaO content in cement, 0.65 as used by the previous report if not specified; $\alpha_H$ is a degree of hydration; $M_{CaO}$ is molar mass of CaO and equal to 56 g/mol and $M_{CO₂}$ is molar mass of CO₂ equal to 44 g/mol. The age factor for microclimatic conditions ($n_m$) associated with the frequency of wetting and drying cycles is $n_m=0$ for sheltered outdoor and $n_m=0.12$ for unsheltered outdoor. $k_{urban}$ is the CO₂ urbanisation factor; and $f_T(t)$ is the temperature factor.

Table 2-1 Ageing factor for CO₂ diffusion coefficient at given water/cement ratio (Yoon et al 2007).

<table>
<thead>
<tr>
<th>w/c</th>
<th>$n_d$</th>
</tr>
</thead>
<tbody>
<tr>
<td>0.45</td>
<td>0.218</td>
</tr>
<tr>
<td>0.50</td>
<td>0.235</td>
</tr>
<tr>
<td>0.55</td>
<td>0.240</td>
</tr>
</tbody>
</table>

For an existing concrete structure, Eqn. (2-1) can be calibrated by carbonation depth measurements. If carbonation depth $x_{test}$ (cm) is known at time of test $t_{test}$, and the structure was constructed in the year $t_{const}$, then the equivalent first-year diffusion coefficient $D_1$ is

$$D_1 = \frac{a}{2f_T(t_{test})} \left( t_{test} - t_{const} \right)^{n_d} k_{urban} \int_{t_{test}}^{t_{const}} C_{CO₂}(t)dt$$

$$f_T(t) = e^{\frac{E}{R} \left( \frac{1}{273+t_{av}(t)} \right)}$$

where $T_{av}(t) = \frac{\sum_t T(t)}{t-t_{const}}$ (2-4)

where $T(t)$ is the temperature in °C, $f_T(t)$ is the temperature correction factor for diffusion coefficient, $E=40$ kJ/mol and $R = 8.314\times10^{-3}$ kJ/mol K.

The CO₂ concentration in 1960 was $\mu_{CO₂}(t) = 316$ ppm, and is assumed to increase linearly to $\mu_{CO₂}(t) = 369$ ppm by 2000. If time of test is before 1960 then CO₂ concentration is assumed constant at $\mu_{CO₂}(t) = 316.0$ ppm for all years to 1960. The variability of CO₂ concentration to year 2000 is zero, i.e., $\sigma_{CO₂}(t) = 0$.

---

1 The carbon dioxide concentration in 1960 and 2000 is assumed on the basis of the information from record derived from spline smoothing of the DE08 and DE08-2 ice cores by Etheridge et al of CSIRO and CSIRO GASLAB Flask Sampling Network, which are available at Carbon Dioxide Information Analysis Centre (http://cdiac.ornl.gov).
Based on the information from Bureau of Meteorology Australia, the median temperature increase due to global warming prior to 2000 will be less than after 2000. For example, from 1950 to 2000 the median temperature increase was only 1.15 °C for Sydney. Simulation analyses have found that the best fit carbonation depth and chloride concentration are not significantly affected by pre-2000 temperature history (less than 1% in most cases). Hence, it is assumed that the mid global temperature change is $T_{\text{mid-test}}=-1.0\,^\circ\text{C}$ in 1950 and temperature increases linearly until $T_{\text{mid-test}}=0\,^\circ\text{C}$ in 2000. If time of test is before 1950 then $T_{\text{mid-test}}$ is assumed constant at $T_{\text{mid-test}}=-1.0\,^\circ\text{C}$ for all years to 1950. The site-specific correction factor for temperature increase ($k_{\text{mid}}$) is omitted due to uncertainty about $T_{\text{mid-test}}$ prior to 2000. The mid temperature is $T(t)=T_{\text{ref}}+T_{\text{mid-test}}(t)$ where $T_{\text{ref}}$ is the reference temperature at year 2000 and $t<2000$. If $t>2000$, then $T(t)=T_{\text{ref}}+[k_{\text{mid}}\times T_{\text{mid}}(t)]$ where $k_{\text{mid}}$ is the site-specific correction factor for temperature increase supplied by the user, and $T_{\text{mid}}$ is the mid global temperature increase for the IPCC emission scenarios.

As known, carbonation is highest for RH=40-75%, or 50-70% (Russell et al 2001). Al-Khayat and Fattuhi (2002) report that little or no carbonation occurs below a RH of 30%, whereas Russell et al (2001) state that below 50% RH there is insufficient moisture for carbonation reactions to take place. Most carbonation models assume RH>50%. To be conservative, assume that if RH(t) < 40% then carbonation front ceases to advance, or carbonation depth does not increase. Relative humidity prior to 2000 is likely to be well over 40% for most infrastructure locations in Australia, otherwise there would be no carbonation and no corrosion - and so no test results to analyse. Hence, relative humidity is not considered in the development of diffusion coefficient from test data.

Carbonation depths obtained using a phenolphthalein pH indicator are measured to an accuracy of ±0.5 mm (e.g., Al-Khayat et al. 2002, Jones et al. 2000). It follows that the diffusion coefficient $D_1$ is variable. If the test data describes a single carbonation depth $x_{\text{test}}$ then we assume that $D_1$ is normally distributed, and there is a 95% probability that $D_1$ lies between those calculated for $x_{\text{test}}\pm0.5$ mm, then the standard deviation of $D_1$ ($\sigma_{D_1}$) is $[D_1(x_{\text{test}}+0.5\,\text{mm}) - D_1(x_{\text{test}}-0.5\,\text{mm})]/3.92$. The mean of $D_1$ is based on diffusion coefficient calculated using $x_{\text{test}}$ given by Eqn. (2-3).

If the test data describes a range of carbonation depths $x_{\text{test}(1)}$ and $x_{\text{test}(2)}$ then we assume that mean($D_1$) is based on $x_{\text{test}}=(x_{\text{test}(1)}+x_{\text{test}(2)})/2$ and if $D_1$ is normally distributed, and there is a 95% probability that $D_1$ lies between those calculated for $x_{\text{test}(1)}$ and $x_{\text{test}(2)}$, then $\sigma_{D_1}= [D_1(x_{\text{test}(2)}) - D_1(x_{\text{test}(1)})]/3.92$.

For times less than $t_{\text{test}}$ it is assumed that $\sigma_{D_1}=0$ and Coefficient of Variation (COV) of $k_{\text{urban}}$ is zero. For all times COV for $n_d$ is zero as the best fit for diffusion coefficient will minimise any carbonation depth model errors and so minimise parameter variability.

It is recognised that carbonation test methods may result in variability in test measurements (e.g., Jones et al. 2000). However, test measurements based on averaging several sample results should minimise these uncertainties. The present analysis
assumes that the carbonation depth obtained using a phenolphthalein pH indicator is an accurate indicator of carbonation.

The carbonation depth for times greater than \( t_{\text{test}} \) is predicted as:

\[
\begin{align*}
\Delta_c(t) &\approx \sqrt{\frac{2f_T(t)D_{CO_2}(t)}{a}} \int_{t_{\text{test}}}^{t} k_{\text{urban}} \left( C_{CO_2}(t) \right) \left( \frac{1}{t - t_{\text{const}}} \right)^{n_a} dt \\
&\quad \quad \quad \quad \left( t \geq t_{\text{test}} \geq 2000 \right) \quad (2-5)
\end{align*}
\]

where \( a, \alpha_H \), and \( f_T(t) \) are given by Eqns. (2-2) and (2-4).

It is observed that corrosion may occur when the distance between the carbonation front and the reinforcement bar surface is less than 1-5 mm (e.g., Yoon et al 2007). However, probabilistic analyses for assessing durability design specifications tend to ignore this effect (Duracrete 2000b, fib 2006). Hence, time to corrosion initiation \( (T_i) \) occurs when carbonation front equals concrete cover, and the probability of corrosion initiation is given by:

\[
\begin{align*}
\Pr[h - \Delta_c(t') < 0] &= p_i(t') \\
&\quad \quad \quad \quad \left( 2-7 \right)
\end{align*}
\]

where \( h \) is the cover thickness.

2.2 Chloride-Induced Corrosion Modelling with Calibration by Field Testing Data

There is evidence to suggest that chloride action is accelerated by carbonation (and \( SO_2 \), \( NO_x \)) because carbonation disturbs the equilibrium between free and bound chlorides in the concrete, thereby increasing the free chloride concentration in the pore solution. However, it appears that this evidence has not been translated into any useful quantitative models. Thus the interaction effect between carbonation and chlorides is, for the time being, omitted from the present study. Use total (acid soluble) chloride content.

The British Standard BS 1881-Part 124 (1988) “Testing concrete - Part 124: Methods for Analysis of Hardened Concrete” is used for assessing total chloride content. Within the U.K. this is seen as the ‘reference’ method (Bamforth et al. 1997). While the test method is accurate, there is little information about variability of test results from ‘identical’ specimens.

Chloride content \( C(x,t) \) is obtained for specimens typically of 10 to 20 mm depth, and the mid-height value of each slice \( (x) \) is used when determining chloride profiles at time \( t \) (e.g., McGee 1999, Tamimi et al 2008).

Chloride penetration based on Fick’s law is
where $C_o$ is the surface chloride concentration, $C_i$ is the initial chloride concentration that already exists in the concrete mix, $D_c$ is the chloride diffusion coefficient, and $x$ is depth of concrete specimen.

Most durability studies, including Duracrete (2000a) assume that initial chloride concentration $C_i$ is zero. However, Polder and de Rooij (2005) and Kershel (2009) show that chloride measurements in the tail of the chloride profile may be considerably higher and for a better fitting result $C_i$ is non-zero. This is due to small, though often acceptable, chloride concentrations in mixing water and aggregates. For example, Kershel (2009) found that chloride content taken at 120 mm from the surface of a bridge in Ireland had chloride levels of 0.011% (by mass of concrete), and Polder and de Rooij (2005) reported values of $C_i$= 0.01%. The effect of $C_i$ on curve fitting will be illustrated later in this Section.

The diffusion coefficient is time-dependent due to the continuing densification of the concrete microstructure as a result of the continuing hydration of the cement. Hence, the following power function is used (eg. Duracrete 2000a):

\[
D_c(t) = D_{c-o} \left( \frac{t_o}{t - t_{const}} \right)^n = D_{c-o} \left( \frac{t_{test} - t_{const}}{t_{test} - t_{const}} \right)^n \quad t \geq t_{test} \geq 2000
\]  

(2-9)

where $D_{c-o}$ is the reference diffusion coefficient at a reference time $t_o$ (age of structure at time of test) and $n$ is the age reduction factor which depends mainly on mix proportions of the concrete. The age reduction factor ($n$) is approximately 0.2, 0.4 and 0.6 for OPC, PFA and GGUBS concrete, respectively (Kershel 2009).

Non-linear least squares methods are used to determine the ‘best fit’ Fick’s law parameters for diffusion coefficient $D_{c-o}$ and surface chloride concentration $C_o$ from Eqns. (2-8) and (2-9). Considering $t=t_{test}$ and time $t_o$ is equal to the age of structures at time of test ($t_{test} - t_{const}$), Eqn. (2-8) can be rewritten as:

\[
C(x, t_{test}) = C_o - (C_o - C_i) \text{erf} \left( \frac{x}{2\sqrt{f_T(t_{test})D_{c-o}(t_{test} - t_{const})}} \right) \quad t_{test} \geq 2000
\]  

(2-10)

where $f_T(t)$ is given by Eqn. (2-4). Note that environment, curing and test factors described in Duracrete (2000a) are assumed as unity for the purposes of best-fit calculations.

As a result, chloride concentration can then be predicted at any time $t$ as:
The 95% confidence intervals to the best fit profile and regression coefficient \( R^2 \) are obtained from standard statistical methods (Ang and Tang 2007). The value of initial chloride concentration \( (C_i) \) can be varied by the user to improve the fit by maximising the regression coefficient. As part of the least squares regression analysis, the user is required to input initial (guess) parameters for \( D_{c-o} \) and \( C_o \) - suggested initial values are \( D_{c-o}=0.5 \) and \( C_o=\text{highest sample concentration} \). If the best fit parameters fail to converge then other initial values should be used until convergence occurs.

The variability of chloride concentration will be derived from the 95% confidence intervals to the best fit profile. If it is assumed that chloride concentration variability is represented as a normal distribution, then the 95% confidence intervals represent a deviation from the mean of \( 1.96 \sigma_C \) where \( \sigma_C \) is the standard deviation of chloride concentration. In addition, in all cases the age factor \( (n) \) is kept at a constant value \( (n=0.2 \text{ for OPC}) \) for simplicity in simulation, though it may only be taken as approximation for other types of concrete.

It is expected that chloride concentration will reduce as depth increases. However, chloride concentrations in the top few mm of the concrete cover may be lower because the concrete skin has a different composition compared to the internal concrete. This “skin effect” can be due to contact with formwork, segregation of aggregates, dielectric reactions between the concrete surface and chloride environment, or chlorides at the surface may be washed out in preparation of sampling (e.g., Andrade et al 1997, Song et al. 2008). If the chloride concentration is lower for the lowest sample depth then this indicates a “skin effect” and this measurement should be omitted from the best fit statistical analysis.

In addition, the statistics of critical chloride concentration \( (C_{cr}) \) are mean = 3.35 kg/m\(^3\), COV=0.375, truncated at 0.35 kg/m\(^3\) - normal distribution (Val and Stewart 2003). These statistics are not affected by concrete quality (Duracrete 2000a). In this regard, the time to corrosion initiation \( (T_i) \) occurs when chloride concentration exceeds critical chloride concentration \( (C_{cr}) \), and the probability of chloride-induced corrosion is:

\[
p_i(t') = \Pr[C(h,t)-Cr < 0] \quad (2-12)
\]

### 2.3 Cover Measurement

Data analysis of carbonation and chloride levels often also includes cover measurements. A covermeter is typically used for these measurements. The British Standard BS1881-Part 204 (1988) “Testing Concrete: Recommendations on the Use of
Electromagnetic Covermeters” states that the indicated cover to steel reinforcement by a calibrated covermeter should be accurate to within ±5% or ±2 mm, whichever is the greater, over the working range given by the manufacturer. A study by Barnes and Zheng (2008) found that all the measured covers were well within ranges specified by BS1881. This study will assume the same accuracy levels for covermeters as BS1881-Part 204 (1988). Hence, assuming a normal distribution and that 95% of covers lie within covermeter accuracy, then the standard deviation is the greater of 2/1.96 mm or 0.05×cover/1.96.

2.4 Modelling of Adaptations for Existing Structures

An adaptation may influence carbonation depth, chloride diffusion coefficient, chloride concentration, corrosion rate, or critical chloride concentration if the existing concrete cover is not replaced. If the existing concrete cover is replaced with new concrete then in this case the deterioration process of the existing structure ceases and the deterioration process restarts at the time of cover replacement at time $t_{adapt}$.

It assumes that any coatings are adequately maintained during the life of the structure (ie. coating will maintain the performance with time). In most cases, an adaptation can be directly simulated by the estimation of correction (multiplicative) factors ($R$) that influence carbonation depth, diffusion coefficient, chloride concentration, critical chloride concentration or corrosion rate.

2.4.1 Carbonation Depth Factor ($R_{\text{carb}}$)

Reduction in carbonation depth only applies when $t$ exceeds $t_{adapt}$, hence the carbonation depth multiplicative factor ($R_{\text{carb}}$) applies only to the incremental increase in carbonation depth each year, namely, carbonation depth is

$$x_c(t) = \begin{cases} x_c(t) & t < t_{adapt} \\ x_c(t-1) + R_{\text{carb}} [x_{c(1)}(t) - x_{c(1)}(t-1)] & t \geq t_{adapt} \end{cases}$$

(2-13)

where $x_{c(1)}$ is the carbonation depth assuming no adaptation measures ($R_{\text{carb}}=1.0$). For example, a carbonation depth multiplicative factor ($R_{\text{carb}}$) of 0.5 to carbonation depth means that calculated corrosion depths are multiplied by 0.5.

Researches indicated that an acrylic-based surface coating causes 10-65% reduction in carbonation based on nine week tests (Ho and Harrison 1990). If a minimum of two coats were applied then the minimum reduction in carbonation depth is 25% (Ho and Harrison 1990). It is reported that there was 60-83% reduction in carbonation depth on the basis of a 2.5 year test (Swamy et al. 1998). Moreno et al (2007) found an 85% reduction in carbonation depth, 64 days after acrylic and 'good quality' vinyl-acrylic coatings, but indicated that most vinyl-acrylic coatings had smaller reductions in
carbonation depth. In addition, it is found that silicone coating has no reduction effect on carbonation depth (Ho and Harrison 1990).

In general, acrylic-based surface coating appears most effective in dry environments, and should not be used in wet environments (Exposures C, C1 and C2). Meanwhile, treatment should be uniformly applied with the minimum number of defects. In the simulation of coatings for existing concrete structures, it is assume that the reduction in carbonation depths is in the range of 25%-75%, or \( R_{\text{carbon}} = 0.25-0.75 \).

### 2.4.2 Chloride Diffusion Factor \((R_D)\)

Reduction in chloride diffusion coefficient only applies when \( t > t_{\text{adapt}} \), hence the chloride diffusion multiplicative factor \((R_D)\) applies only to the incremental increase in chloride concentration each year, namely, chloride concentration at time \( t \) is

\[
C(x,t) = C_o - (C_o - C_i) \text{erf} \left( \frac{x}{2 \int f_T(t) D_{c-o} \left( \frac{t_{\text{test}} - t_{\text{const}}}{t - t_{\text{const}}} \right)^{0.5} (t - t_{\text{const}})} \right)
\]

\( t < t_{\text{adapt}} \)  

\[
C(x,t) \approx C_o - (C_o - C_i) \text{erf} \left( \frac{x}{2 \int f_T(t) \left( \frac{t_{\text{test}} - t_{\text{const}}}{t - t_{\text{const}}} \right)^{0.5} [D_{c-o} (t_{\text{adapt}} - t_{\text{const}}) - 1] + R_D D_{c-o} (t - t_{\text{adapt}} + 1)] \right)
\]

\( t \geq t_{\text{adapt}} \)

where \( D_{c-o} \) is the chloride diffusion coefficient assuming \( R_D = 1.0 \). Eqn. (2-14) is an approximation, and aims to reflect the time integration of diffusion coefficient within the error function.

Surface treatments aim to reduce the chloride diffusion coefficient: and they should be applied before corrosion initiation. The effective diffusion coefficient considering surface coating is described by (Buenfeld and Zhang 1998):

\[
D_e = \frac{T_c + T_{\text{ST}}}{\frac{T_c}{D_c} + \frac{T_{\text{ST}}}{D_{\text{ST}}}}
\]

\( T_c = \) concrete cover  
\( T_{\text{ST}} = \) thickness of surface treatment  
\( D_c = \) chloride diffusion coefficient without surface treatment  
\( D_{\text{ST}} = \) chloride diffusion coefficient of surface treatment
Table 2-2 Parameters of surface coatings

<table>
<thead>
<tr>
<th>Surface Treatment</th>
<th>$T_{ST}$ (mm)</th>
<th>$D_{ST}$ ($10^{-12}$ m$^2$/s)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Polyurethane sealer</td>
<td>0.02</td>
<td>$1.4\times10^{-1}$</td>
</tr>
<tr>
<td>Silane</td>
<td>2</td>
<td>$2.5\times10^{-1}$</td>
</tr>
<tr>
<td>P-m (Polymer modified) cementitious coating</td>
<td>1.5</td>
<td>$6.3\times10^{-3}$</td>
</tr>
</tbody>
</table>

Note: $D_c=3\times10^{-12}$ for tests used to determine $D_{ST}$.

An acrylic coating reduces diffusion coefficient by 50% ($R_D=0.5$), and a two component epoxy resin reduced diffusion coefficient by 100% ($R_D=0.0$) (Aguair et al. 2008).

### 2.4.3 Chloride Concentration Factor ($R_{chloride}$)

Reduction in chloride concentration only applies when $t$ exceeds $t_{adapt}$, hence the chloride concentration multiplicative factor ($R_{chloride}$) applies only to the incremental increase in chloride concentration each year, namely, chloride concentration is:

$$C(x,t) = \begin{cases} 
C(x,t)_{(1)} & t < t_{adapt} \\
C(x,t-1) + R_{chloride} \left[ C(x,t)_{(1)} - C(x,t-1)_{(1)} \right] & t \geq t_{adapt}
\end{cases}$$

(2-16)

where $C(x,t)_{(1)}$ is the chloride concentration assuming no adaptation measures ($R_{chloride}=1.0$).

Electrochemical Chloride Extraction removes some chlorides from concrete cover. Sharp et al (2002) in a review of U.S. chloride extraction studies (bridge decks and piers) found that chloride extraction removed 20-90% of chlorides. Schneck et al. (2007) found a 70% reduction in chloride content, and Velivavaskis et al. (1998) reported 40-80% reduction from case studies taken from U.S. bridges. An experimental study found that 33-76% of chlorides were removed (Sharp et al 2002). Carbonation will reduce chloride extraction rate (Ihekwaba et al 1996) and so a lower limit of 30% extraction may be appropriate ($R_{chloride}=0.7$).

### 2.4.4 Critical Chloride Concentration Factor ($R_{cr}$)

The critical chloride concentration multiplicative factor ($R_{cr}$) is applied to the critical chloride concentration ($C_{cr}$) as:

$$C_{cr} = \begin{cases} 
C_{cr} & t < t_{adapt} \\
R_{cr}C_{cr} & t \geq t_{adapt}
\end{cases}$$

(2-17)
Corrosion Rate Factor ($R_{icorr}$)

The corrosion rate multiplication factor ($R_{icorr}$) is applied to the corrosion rate at $T=20^\circ C$ as

$$i_{corr-20} = i_{corr-20} \quad t < t_{adapt}$$

$$i_{corr-20} = R_{icorr}i_{corr-20} \quad t \geq t_{adapt}$$

(2-18)

The corrosion rate factor is applied to simulate realkanisation. Realkanization is an electrochemical process to raise pH near reinforcement. After realkanization then concrete will not easily recarbonate and so the effect tends to be permanent (Velivavaskis et al. 1998). Yeih and Chang (2005) report that the corrosion rate becomes negligible ($<0.1 \mu A/cm^2$) after realkanization. Hence, in the simulation, the corrosion rate factor $R_{icorr}=0.0$.

Surface coatings may also reduce corrosion rate because of reduced moisture ingress into concrete (Bentur et al. 1997; Ibrahim et al. 1999). However, it will be assumed no effect on corrosion rate in the simulation, i.e. $R_{icorr}=1.0$.

Cathodic Protection is assumed to be 100% effective in removing chloride ions from reinforcement, and therefore result in $R_D=0.0$ and $R_{icorr}=0.0$.

Chloride extraction will lower corrosion rate by 60-90% (Velivavaskis et al. 1998), and Broomfield (1997) states 'very low corrosion rates' even if chloride levels are above the critical chloride content. It is therefore assumed that $R_{icorr}=0.5$ for the simulation.

2.4.5 Replace Existing Cover with New Concrete

The existing concrete cover is removed and replaced with new concrete. In this case, the deterioration process of the existing structure ceases and the deterioration process restarts at the time of cover replacement at time $t_{adapt}$. It is assumed that (i) the new concrete cover re-passivates the steel reinforcement (corrosion rate is zero), (ii) corrosion loss of reinforcement is not rectified at time of cover replacement, and (iii) if corrosion initiates the corrosion rate is the same as for the original concrete cover (i.e. not affected by the new concrete cover).

The carbonation depth at time $t$ is:

$$x_c(t) \approx \sqrt{\frac{2f_{TA}(t)D_{CO_2}(t)}{a} k_{urban} \int_{t_{adapt}}^{t} C_{CO_2}(t)dt \left(\frac{1}{t - t_{adapt}}\right)^{n_m} \left(1 - e^{-3.38w/c}\right) t \geq t_{adapt} \geq 2000}$$

$$D_{CO_2}(t) = D_i \left(t - t_{adapt}\right)^{-n_d} \quad a = 0.75C_eCaO\alpha_{M} \frac{M_{CO_2}}{M_{CaO}} \quad \alpha_{M} \approx 1 - e^{-3.38w/c}$$

(2-20)
\[ f_{TA}(t) \approx e^{\left( \frac{E}{R} \left( \frac{1}{273 + T_{av} - A(t)} \right) \right)} \]

where \( T_{av-A}(t) = \frac{t_{\text{cur}}}{t - t_{\text{adapt}}} \)

(2-21)

where parameters w/c, \( C_e \), \( D_1 \) and \( n_d \) are suggested in Table 2. The standard deviation for \( D_1 \) is approximately 0.15, and COV for \( n_d \) is approximately 0.12 for all w/c ratios. These statistics represent model error (or accuracy). However, the user may select statistical parameters according to their own experience.

<table>
<thead>
<tr>
<th>( F'c ) (MPa)</th>
<th>w/c ( d ) ratio</th>
<th>( C_e^a ) (kg/m(^3))</th>
<th>mean ( D_1 ) ((\times 10^{-3} \text{cm}^2\text{s}^{-1}))</th>
<th>mean ( n_d )</th>
</tr>
</thead>
<tbody>
<tr>
<td>20</td>
<td>0.56</td>
<td>320</td>
<td>2.22</td>
<td>0.240</td>
</tr>
<tr>
<td>25</td>
<td>0.56</td>
<td>320</td>
<td>2.22</td>
<td>0.240</td>
</tr>
<tr>
<td>32</td>
<td>0.50</td>
<td>320</td>
<td>1.24</td>
<td>0.235</td>
</tr>
<tr>
<td>40</td>
<td>0.46</td>
<td>370</td>
<td>0.65</td>
<td>0.218</td>
</tr>
<tr>
<td>50</td>
<td>0.40</td>
<td>420</td>
<td>0.47(^b)</td>
<td>0.19(^c)</td>
</tr>
</tbody>
</table>

Notes - a: based on minimum cement content specified in AS 5100.5, b: conservative value, c: conservative value, \( d \): based on maximum w/c ratio specified in AS 5100.5.

The chloride concentration at time \( t \) is (Duracrete 2000a):

\[
C(x,t) = C_o \left[ 1 - \text{erf} \left( \frac{x}{2 \sqrt{k_e \cdot f_{TA}(t)D_c \left( \frac{t_o}{t - t_{\text{adapt}}} \right)^n \cdot (t - t_{\text{adapt}}) \right)} \right) \right] \quad t \geq t_{\text{adapt}} \geq 2000
\]

(2-22)

where \( t_o \) is 28 days (0.0767 years), \( f_{TA} \) is given by Eqn. (18), and \( D_c \), \( n \), \( k_e \) and \( C_o \) are given in Tables 3 and 4. Since statistical parameters given in Table 3 are sourced from Duracrete (2000a) which is based on Fick's Law with \( C_i=0.0\% \), then the same formulation of Fick's Law is used herein. However, the user may select statistical parameters according to their own experience.

Table 2-6 and Table 2-7 summarise the factors selected for the simulation of various adaptation measures described previously, which are in fact the traditional maintenance techniques, for carbonation- and chloride-induced corrosion.
### Table 2-4  Input Parameters for New Concrete Cover, for Chlorides

<table>
<thead>
<tr>
<th>Exposure Descriptions</th>
<th>Exposure Classification</th>
<th>$F'_c$ (MPa)</th>
<th>w/c ratio</th>
<th>mean($D_c$) $\times 10^{12}$ COV=0.285</th>
<th>mean(n) $\sigma$=0.07</th>
<th>mean(ke) SD</th>
</tr>
</thead>
<tbody>
<tr>
<td>Members in exterior environments</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Near-coastal (1-50 km, any climate)</td>
<td>B1</td>
<td>32</td>
<td>0.50</td>
<td>15</td>
<td>0.65</td>
<td>0.676 $\sigma$=0.114</td>
</tr>
<tr>
<td>Coastal (up to 1 km, excluding tidal and splash zones), any climate</td>
<td>B2</td>
<td>40</td>
<td>0.46</td>
<td>10</td>
<td>0.65</td>
<td>0.676 $\sigma$=0.114</td>
</tr>
<tr>
<td>Surfaces of members in water</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>splash and tidal zones</td>
<td>C</td>
<td>50</td>
<td>0.40</td>
<td>7</td>
<td>0.37</td>
<td>0.924 $\sigma$=0.155</td>
</tr>
<tr>
<td>in spray zone (&gt;1 m above wave crest level)</td>
<td>C1e</td>
<td>50</td>
<td>0.40</td>
<td>7</td>
<td>0.37</td>
<td>0.265 $\sigma$=0.045</td>
</tr>
<tr>
<td>splash and tidal zones</td>
<td>C2e</td>
<td>50</td>
<td>0.40</td>
<td>7</td>
<td>0.37</td>
<td>0.924 $\sigma$=0.155</td>
</tr>
</tbody>
</table>

Notes - b: based on maximum w/c ratio specified in AS 5100.5, e: New Concrete Structures Code - AS3600-2009 (based on draft D05252 (2005)).

### Table 2-5  Surface Chloride Concentration $C_0$ (Val and Stewart 2003).

<table>
<thead>
<tr>
<th>Environment Exposure Descriptions</th>
<th>Mean</th>
<th>COV</th>
<th>Distribution</th>
</tr>
</thead>
<tbody>
<tr>
<td>Splash/Tidal Zone</td>
<td>7.35 kg/m$^3$</td>
<td>0.7</td>
<td>Lognormal</td>
</tr>
<tr>
<td>Atmospheric Zone on the Coast</td>
<td>2.95 kg/m$^3$</td>
<td>0.7</td>
<td>Lognormal</td>
</tr>
<tr>
<td>Atmospheric Zone &gt;1 km from the Coast</td>
<td>1.15 kg/m$^3$</td>
<td>0.5</td>
<td>Lognormal</td>
</tr>
</tbody>
</table>

### Table 2-6  Selection of various factors of adaptation measures for carbonation-induced corrosion

<table>
<thead>
<tr>
<th>Adaptation measures</th>
<th>carbonation depth factor $[R_{carb}]$</th>
<th>carbonation diffusion factor $[R_D]$</th>
<th>corrosion rate factor $[R_{corr}]$</th>
</tr>
</thead>
<tbody>
<tr>
<td>acrylic-based surface coating</td>
<td>0.25-0.75</td>
<td>1</td>
<td>1</td>
</tr>
<tr>
<td>realkanization</td>
<td>1</td>
<td>1</td>
<td>0</td>
</tr>
<tr>
<td>replace existing cover</td>
<td>1</td>
<td>1</td>
<td>1</td>
</tr>
</tbody>
</table>
### Table 2-7 Selection of various factors of adaptation measures for chloride-induced corrosion

<table>
<thead>
<tr>
<th>Adaptation measures</th>
<th>Chloride concentration factor [R&lt;sub&gt;chloride&lt;/sub&gt;]</th>
<th>Chloride diffusion factor [R&lt;sub&gt;D&lt;/sub&gt;]</th>
<th>Critical chloride concentration factor [R&lt;sub&gt;cr&lt;/sub&gt;]</th>
<th>Corrosion rate factor [R&lt;sub&gt;icorr&lt;/sub&gt;]</th>
</tr>
</thead>
<tbody>
<tr>
<td>acrylic coating</td>
<td>1</td>
<td>0.5</td>
<td>1</td>
<td>1</td>
</tr>
<tr>
<td>epoxy resin coating</td>
<td>1</td>
<td>0</td>
<td>1</td>
<td>1</td>
</tr>
<tr>
<td>cathodic protection and concrete</td>
<td>1</td>
<td>0</td>
<td>1</td>
<td>0</td>
</tr>
<tr>
<td>replacement</td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>electrochemical chloride extraction</td>
<td>0.7</td>
<td>1</td>
<td>1</td>
<td>0.5</td>
</tr>
<tr>
<td>replace existing cover</td>
<td>1</td>
<td>1</td>
<td>1</td>
<td>1</td>
</tr>
</tbody>
</table>
3. CLIMATE CHANGE IMPACT ON CORROSION OF EXISTING CONCRETE BRIDGE STRUCTURES IN TEMPERATE CLIMATE ZONES

3.1 Chloride-Induced Corrosion of Concrete Bridges

3.1.1 Bridges for Chloride-Induced Corrosion Assessment

Eleven existing bridges located within temperate climate zones in NSW are used as case studies of climate change impact on chloride-induced corrosion of the bridges. The bridges were chosen as representative of different construction periods at different regions, as shown in Table 3-1. The construction periods of bridges are divided into groups of prior to 1959, 1959-1970, 1971-1994, and post-1995. The data of post-1995 bridge structures is not available.

The profiles of chloride concentration were tested in 2008 at various locations of the bridges, and the test results are reported by (RTA, 2008). In this study, analysis will be carried out for the location where the core sample shows the highest chloride concentration among all the test samples of the each bridge at the same exposure. For example, as shown in Table 3-1, the selected samples for the bridges are given by in the report (RTA, 2008), and C1 is the exposure class of the sample. Exposure C1 is for locations more than 1m above the water level or spray zones, and exposure C2 is for locations less than 1m above the water level or splash and tidal zones.
The field test data is then applied to calibrate the diffusion coefficient and surface chloride concentration, as indicated in the previous chapter, which are then used to project the corrosion until 2100 with the effect of climate change or not, as well as the influence of various adaptation measures.

Table 3-1 List of bridges assessed for corrosion due to chloride penetration

<table>
<thead>
<tr>
<th>Construction Period</th>
<th>Southern Region</th>
<th>Sydney Region</th>
<th>Hunter Region</th>
<th>Northern Region</th>
</tr>
</thead>
<tbody>
<tr>
<td>Pre-1959</td>
<td>Bridge BA1</td>
<td>Bridge BB1</td>
<td>Bridge BC1</td>
<td>Bridge BD1</td>
</tr>
<tr>
<td></td>
<td>Built year: 1956</td>
<td>Built year: 1925</td>
<td>Built year: 1940</td>
<td>Built year: 1943</td>
</tr>
<tr>
<td></td>
<td>Exposure: C1</td>
<td>Exposure: C1</td>
<td>Exposure: C1</td>
<td>Exposure: C2</td>
</tr>
<tr>
<td></td>
<td>Cover: 55mm</td>
<td>Cover: 29mm</td>
<td>Sample: M1</td>
<td>Cover: 70mm</td>
</tr>
<tr>
<td></td>
<td></td>
<td></td>
<td>(C2)</td>
<td></td>
</tr>
<tr>
<td></td>
<td></td>
<td></td>
<td>Cover: 50mm</td>
<td></td>
</tr>
<tr>
<td>1959 – 1970</td>
<td>Bridge BA2</td>
<td>Bridge BB2</td>
<td>Bridge BC2</td>
<td>Bridge BD2</td>
</tr>
<tr>
<td></td>
<td>Exposure: C2</td>
<td>Exposure: C1</td>
<td>Exposure: C2</td>
<td>Exposure: C2</td>
</tr>
<tr>
<td></td>
<td>Cover: 50mm</td>
<td>Cover: 48mm</td>
<td>Cover: 53mm</td>
<td>Cover: 62mm</td>
</tr>
<tr>
<td>1971 – 1994</td>
<td>Bridge BA3</td>
<td>Bridge BB3</td>
<td>Bridge BC3</td>
<td>Bridge BD3</td>
</tr>
<tr>
<td></td>
<td>Exposure: C2</td>
<td></td>
<td>Exposure: C2</td>
<td>Exposure: C1/C2</td>
</tr>
<tr>
<td></td>
<td>Cover: 58mm</td>
<td></td>
<td>Cover: 67mm</td>
<td>Cover: 62mm</td>
</tr>
<tr>
<td>Post-1995</td>
<td>Test data not available</td>
<td>Test data not available</td>
<td>Test data not available</td>
<td>Test data not available</td>
</tr>
</tbody>
</table>

3.1.2 Pre-1959 Bridge Structures

Bridge BA1 (1956, Southern Region)

Bridge BA1, as shown Figure 3-1, is constructed in the pre-1959 period and by 2008 has been in service for 52 years with fairly extensive corrosion and damage at the inspected location under exposure C1.
The probability of corrosion initiation was 48% by 2000, and the corresponding probability of corrosion damage was 36% and the mean rebar loss was around 0.55 mm. They are now estimated to be 62%, 52% and 0.97 mm, respectively. If left untreated, the probability of corrosion initiation and damage increases to above 97% by 2100 in all three climate change scenarios. Thus it is almost certain that the bridge will experience damage, as shown in Figure 3-2. The effect of climate change on the probability is within 1%, and the effect on rebar loss within 1.2 mm by 2100 (13% of the mean rebar loss in the absence of climate change). In fact, the very severe chloride-induced corrosion on this bridge structure has subdued the effect of changing climate. It would be more worthwhile to assess the impact of adaptation measures that are applied to ameliorate the deterioration caused by the corrosion.
Corrosion initiation probability

Corrosion damage probability

Mean rebar loss (mm)

Figure 3-2 The probability of chloride-induced corrosion initiation and damage and mean rebar loss of Bridge BA1 in a temperate climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘year 2000 level’ is the relevant value in the absence of climate change.

Bridge BB1 (1925, Sydney) and BB2 (1959, Sydney)

Concrete structures on bridges built in 1925 and 1959 were assessed as shown in Figure 3-3. One built in 1959 shows very low corrosion risk at a location with exposure C1 as the probabilities of corrosion and the mean rebar loss were negligibly small at the time of testing in 2008. However, one built in 1925 shows the probability of chloride-induced corrosion initiation and damage at 14% and 12% by 2008, respectively. The corresponding mean rebar loss is 0.25mm.

The probability of corrosion initiation is projected to be 70%, 68% and 67% for A1FI, A1B and 550 ppm stabilisation emission scenarios, respectively, in comparison with 62% in the absence of climate change, as shown in Figure 3-4. The probability of corrosion damage is projected to be 68%, 66% and 65% for the three emission scenario respectively in comparison with 60% in the absence of climate change. The mean rebar loss is 3.7mm, 3.5mm, and 3.4mm for the three emission scenarios respectively in comparison with 2.9mm in the absence of climate change. Therefore, the effect of climate change on the probability of chloride-induced corrosion initiation and damage is within 8% increase in probability value, (or 13% increase in percentage change) and the effect on the mean rebar loss is within 0.8mm increase by 2100, for the three emission scenarios.
Figure 3-3 Bridge BB1 constructed in 1925 and 1959, in Sydney (Source: RTA)

Figure 3-4 The probability of chloride-induced corrosion initiation and damage and mean rebar loss of Bridge BB1(1925) in a temperate climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘year 2000 level’ is the relevant value in the absence of climate change.
Bridges BC1 (1940, Hunter Region)

Bridge BC1 has been in service for 68 years to 2008, as shown in Figure 3-5, the tested structures are exposed to C1 and C2 class, respectively. The structure at a location of the bridge under exposure C1 shows very low corrosion in terms of the negligible probability of corrosion initiation and damage as well as mean rebar loss.

The structure at a location with exposure C2, however, shows considerable likelihood of chloride-induced corrosion. The probabilities of corrosion initiation and damage were very high at 54% and 49%, respectively, and the mean rebar loss is 2.0mm by 2008. As shown in Figure 3-6, the probability of corrosion initiation and damage is above 99%, or almost certain by 2100. The corresponding mean rebar loss is 18mm in comparison with 16mm in the absence of climate change. The significance of the corrosion rate subdues the effect of climate change on corrosion initiation and damage that become almost certain anyway by 2100. It would be more important to consider the impact of climate change on maintenance measures that are applied to reduce the corrosion-induced structural deterioration. However, climate change is projected to cause up to 2.2mm increase in mean rebar loss by 2100, or 14% increase in comparison with the rebar loss in the absence of climate change.
Figure 3-6 The probability of chloride-induced corrosion initiation and damage and mean rebar loss of Bridge BC1 in a temperate climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘Year 2000 level’ is the relevant value in the absence of climate change.

**Bridge BD1 (1943, Northern Region)**

Bridge BD1 has been in service for 65 years to 2008, the structure at a location under exposure C2 was tested and applied to calibrate the corrosion models. It shows moderate corrosion in 2008 with the probabilities of corrosion initiation at 23%, the probability of corrosion damage at 20%, and the mean rebar loss at 1.1mm by 2008. It may be partially contributed by the high cover thickness of 70mm.

As shown in Figure 3-7, the probability of corrosion initiation is projected to increase to 79%, 77% and 76% by 2100 for A1FI, A1B and 550 ppm stabilisation emission scenario in comparison with 73% in the absence of climate change. The probability corrosion damage is 77%, 76%, and 75% by 2100 for the three emission scenarios in comparison with 71% in the absence of climate change. The mean rebar loss is 11.2mm, 10.7mm, and 10.3mm for the three emission scenarios, respectively, in comparison with 9.1mm in the absence of climate change. In this regard, climate change may lead to an increase of up to 6% in probability value of corrosion initiation and damage (or 8% increase in percentage change), and 0.9mm more mean rebar loss by 2100 for the three emission scenarios.
Figure 3-7 The probability of chloride-induced corrosion initiation and damage and mean rebar loss of Bridge BD1 in a temperate climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘year 2000 level’ is the relevant value in the absence of climate change.

3.1.3 1959-1970 Bridge Structures

Bridges BA2 (1959, Southern Region)

Bridge BA2 has been in service for 49 years to 2008 since constructed, as shown in Figure 3-8. The simulation based on the tests of concrete structures at a location with exposure C2 indicates that the probability of corrosion initiation and damage were almost equal to 1.0 or corrosion is almost certain by 2008. The impact of climate change was subdued as it would not be able to make too much difference on the corrosion (or make corrosion more severe). Therefore the climate change impact will not be discussed for this bridge.

Bridge BC2 (1968, Hunter Region)

Bridge BC2 in the Hunter region has been in service for 40 years to 2008 as shown in Figure 3-9, and appears severe corrosion at a location of bridge structures under exposure C2, in terms of the probability of corrosion initiation and damage for more than 90% in 2008. At such a degree of corrosion severity, there is only a very limited impact of climate change.
Figure 3-8 Bridge BA2 constructed in 1959 in the southern region of NSW (Source: RTA)

Figure 3-9 Bridge BC2 constructed in 1968 in Hunter region of NSW (Source: RTA)
As shown in Figure 3-10, by 2100, corrosion initiation and damage are almost certain, regardless of climate change that only causes no more than 0.3% increase in probability value. However, the climate change is projected to cause up to 2.1mm increase of mean rebar loss by 2100, about 10% of the rebar loss in the absence of climate change.

Figure 3-10 The probability of chloride-induced corrosion initiation and damage and mean rebar loss of Bridge BC2 in a temperate climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘year 2000 level’ is the relevant value in the absence of climate change.

Bridge BD2 (1967, Northern Region)

Bridge BD2 in NSW has been in service for 41 years by 2008, as shown in Figure 3-11. The tested part of structures at a location with exposure C2 shows very low corrosion by 2008. It is represented by the low probability of corrosion initiation and damage that was estimated around 3% while the mean rebar loss was negligible.

As shown in Figure 3-12, by 2100, the probability of chloride-induced corrosion initiation is projected to be 46%, 45% and 43% for A1F1, A1B and 550 ppm stabilisation emission scenario, respectively, in comparison with 39% in the absence of climate change. The probability of corrosion damage is projected to be 44%, 43% and 42% for the three emission scenario in comparison with 38% in the absence of climate change. The mean rebar loss is 4.6mm, 4.2mm and 4.0mm by 2100 for the three emission scenarios in comparison 3.3mm in the absence of climate change.
Figure 3-11 Bridge BD2 constructed in 1967 in the northern region of NSW (Source: RTA)

(a) Corrosion initiation probability

(b) Corrosion damage probability

(c) Mean rebar loss (mm)

Figure 3-12 The probability of chloride-induced corrosion initiation and damage and mean rebar loss of Bridge BD2 in a temperate climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘year 2000 level’ is the relevant value in the absence of climate change.
In other words, climate change leads to 7% more in terms of change in probability value of corrosion initiation and damage (or 18% increase in percentage change of corrosion initiation probability due to climate change). It also causes up to 1.2mm increase in mean rebar loss (or 39% more than the mean rebar loss without the effect of climate change).

3.1.4 1971-1994 Bridge Structures

Bridge BA3 (1980, Southern Region)

Bridge BA3 in NSW has been in service for only 28 years by 2008, as shown in Figure 3-13, but structures under exposure C2 at a location shows certain corrosion initiation and corrosion damage with the mean rebar loss of 4.1mm. The effect of climate change on the corrosion initiation and damage of structures is little when the corrosion tends to be certain or almost certain. However, climate change is projected to increase the mean rebar loss by up to 2.8mm, or 14% of the mean rebar loss in the absence of climate change.

Figure 3-13 Bridge BA3 constructed in 1980, in the southern region of NSW (Source: RTA)
Bridge BC3 (1989, Hunter Region)

Bridge BC3 has been in service for 24 years by 2008, showing fairly small extent of corrosion with a negligible mean rebar loss while the probabilities of corrosion initiation is at 3%, and the probability of corrosion damage is about 1%.

However, as shown in Figure 3-15, the probability of corrosion initiation and damage can increase dramatically after 2020, reaching to about 95% in 2060.

The effect of climate change make the probability of corrosion initiation and damage by 4% more than those in the absence of climate change by 2045, and the effect then gets less impact with the increase of the probability or corrosion severity. By 2100, the mean rebar loss is projected to be up to 2.8mm more than the loss in the absence of climate change, or 24% increase.
Figure 3-15 The probability of chloride-induced corrosion initiation and damage and mean rebar loss of Bridge BC3 in a temperate climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘year 2000 level’ is the relevant value in the absence of climate change.

Bridge BD3 (1984, Northern Region)

Bridge BD3 has been in service for 24 years by 2008 since constructed in 1984, as shown in Figure 3-16, and currently shows relatively low likelihood of corrosion at the location of structures under exposure C1 as well at another location of structures under exposure C2. By 2008, the probability of corrosion initiation and damage as well as the mean rebar loss of concrete exposed to C1 were 6%, 2% and negligible mean rebar loss, respectively. For that exposed to C2, the probability was all around 3% by 2008. As shown in Figure 3-17 and Figure 3-18, climate change may have noticeable impact on the probability as well as mean rebar loss.
Figure 3-16 Bridge BD3 constructed in 1984 in the northern region of NSW (Source: RTA)

Figure 3-17 The probability of chloride-induced corrosion initiation and damage and mean rebar loss of concrete structures of Bridge BD3 at exposure C1 in a temperate climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘year 2000 level’ is the relevant value in the absence of climate change.
Figure 3-18 The probability of chloride-induced corrosion initiation and damage and mean rebar loss of concrete structures of Bridge BD3 at exposure C2 in a temperate climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘year 2000 level’ is the relevant value in the absence of climate change.

For the concrete structures of Bridge BD3 at exposure C1 under the effect of climate change, the probability of corrosion initiation is projected to be 72% for A1FI emission scenario, 70% for A1B emission scenario, and 69% for 550 ppm stabilisation emission scenario, in comparison with 65% projected in the absence of climate change by 2100. The probability of corrosion damage is projected to be 69%, 67% and 66% for the three emission scenarios respectively in comparison with 62% projected without the effect of climate change. The mean rebar loss will be 3.5mm, 3.3mm and 3.2mm for the three emission scenarios respectively in comparison with 2.7mm estimated in the absence of climate change. In another word, as a result of the climate change, the mean rebar loss may increase 30%, 22% and 19%, respectively, in association with the emission scenarios.

For the concrete structures of Bridge BD3 at exposure C2 under the effect of climate change, the probability of corrosion initiation is projected to be up to 9% more than the probability projected in the absence of climate change by 2070 (or 13% increase in percentage change). The probability of corrosion damage is up to 10% more by 2075 (or 14% increase in percentage change). The impact of climate change on the
probability value increase is reduced to 5-6% by 2100 (or 6-7% increase in percentage change). In addition, the mean rebar loss will be 10.2mm, 9.6mm and 9.1mm for the three emission scenarios respectively in comparison with 7.5mm by 2100 in the absence of climate change, indicating 36%, 28% and 21% increase in rebar loss. It appears that the impact of climate change is more on concrete at exposure C2 than C1.

3.2 Carbonation-Induced Corrosion of Concrete Bridges

3.2.1 Bridges for Carbonation-Induced Corrosion Assessment
Seven bridges exposed to C1 and C2 in the temperate climate zone in NSW, as shown in Table 3-2, are applied for the assessment of likely impact of climate change. Similar to those bridges for the case study of chloride-induced corrosion, the bridges are divided into four regions, i.e. Southern, Sydney, Hunter and Northern region, and four construction periods, i.e. pre-1959, 1959-1970, 1971-1994 and post-1995. The field carbonation test data was provided with a range of carbonation depth (CR) listed in the Table. However, they are not available for all bridges. The assessment will be done only for those with available carbonation depth information.

Table 3-2. List of bridges assessed for carbonation-induced corrosion

<table>
<thead>
<tr>
<th>Construction Period</th>
<th>Southern</th>
<th>Region</th>
<th>Sydney</th>
<th>Hunter</th>
<th>Northern</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td></td>
<td></td>
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<td></td>
</tr>
<tr>
<td>Pre-1959</td>
<td>Bridge BA1</td>
<td>Bridge BB1</td>
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<td>Test data not available</td>
<td></td>
</tr>
<tr>
<td></td>
<td>1956</td>
<td>1925</td>
<td></td>
<td></td>
<td></td>
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<tr>
<td></td>
<td>CR: 10 – 25 mm</td>
<td>CR: 5 – 35 mm</td>
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<td></td>
<td></td>
</tr>
<tr>
<td></td>
<td>Cover: 55mm</td>
<td>Cover: 29mm</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>1959 – 1970</td>
<td>Bridge BA2</td>
<td>Bridge BB2</td>
<td>Bridge BC2</td>
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<td></td>
</tr>
<tr>
<td></td>
<td>1959</td>
<td>1959</td>
<td>1968</td>
<td></td>
<td></td>
</tr>
<tr>
<td></td>
<td>CR: 1 – 5 mm</td>
<td>CR: 5 – 15 mm</td>
<td>CR: 5 – 25 mm</td>
<td></td>
<td></td>
</tr>
<tr>
<td></td>
<td>Cover: 50mm</td>
<td>Cover: 48mm</td>
<td>Cover: 53mm</td>
<td></td>
<td></td>
</tr>
<tr>
<td>1971 – 1994</td>
<td>Bridge BA3</td>
<td>Test data not available</td>
<td>Test data not available</td>
<td>Bridge BD3</td>
<td></td>
</tr>
<tr>
<td></td>
<td>1980</td>
<td></td>
<td></td>
<td>1984</td>
<td></td>
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<td>Cover: 58mm</td>
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<td>Test data not available</td>
<td>Test data not available</td>
<td>Test data not available</td>
<td></td>
</tr>
</tbody>
</table>
3.2.2 Pre-1959 Bridge Structures

*Bridge BA1 (1956, Southern Region)*

Bridge BA1 was built in 1956. Concrete in the test site has cover of 55mm with exposure C1-C2. The mean carbonation depth in 2008 is 17.5mm as tested. As shown in Figure 3-19, climate change may have noticeable impact on the carbonation depth. By 2100, the mean carbonation depth is 34.0mm for A1FI emission scenario, 31.9mm for A1B emission scenario, and 30.1mm for 550ppm stabilisation scenario in comparison with 26.5mm estimated in the absence of climate change, or 28%, 20% and 14% more, respectively. However, 55mm cover of concrete has almost prevented the occurrence of corrosion initiation and damage represented by their negligible probability along with a very small mean rebar loss, even if climate change would contribute a significant increase in carbonation depth. On the other hand, it demonstrated the importance of proper cover design to reduce the impact of carbonation on corrosion of reinforcement.

![Figure 3-19 Carbonation depth, probability of carbonation-induced corrosion initiation and damage and mean rebar loss of concrete structures of Bridge BA1 in a temperate climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘Year 2000 level’ is the relevant value in the absence of climate change.](image-url)
Bridge BB1 (1925, Sydney)

The part of structures of Bridge BB1 built in 1925 has a concrete cover of 29mm. Field tests found carbonation in the range of 5-35mm, very likely inducing corrosion. As shown in Figure 3-20, the mean carbonation depth is 35.2mm for A1FI emission scenario, 33.3mm for A1B emission scenario, and 31.6 for 550ppm stabilisation emission scenario by 2100, in comparison with 28.4mm estimated without the effect of climate change.

![Figure 3-20](image)

Figure 3-20 Carbonation depth, probability of carbonation-induced corrosion initiation and damage and mean rebar loss of concrete structures of Bridge BB1(1925) in a temperate climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘year 2000 level’ is the relevant value in the absence of climate change.

As carbonation at the mean depth of 28.4mm in the absence of climate change is already very close to cover thickness of 29mm, any increase in carbonation depth caused by climate change may significantly impact on the probability of corrosion initiation and damage. As shown in Figure 3-20, the probability of corrosion initiation is 72%, 67% and 63% for the three emission scenarios respectively in comparison with 51% estimated in the absence of climate change. In other words, the probability increases 21%, 16% and 12% in value, or an increase of 41%, 31% and 14% in percentage change due to climate change. Meanwhile, the probability of corrosion
damage is 63%, 60% and 55% in comparison with 44% estimated in the absence of climate change.

3.2.3 1959-1970 Bridge Structures

Bridge BA2 (1959, Southern Region)

Bridge BA2 was constructed in 1950. The concrete in the test site has 50mm cover. The carbonation depth was identified in the range of 1-5mm, which is much less than the cover thickness. Therefore, the likelihood of carbonation-induced corrosion is very low, and the impact of climate change is unlikely to alter the likelihood substantially though it may affect the carbonation depth as discussed previously.

Bridge BB2 (1959, Sydney)

Bridge BB2, constructed in 1959, has cover of 48mm, and the carbonation depth was tested in the range of 5-15mm that is much less than the cover. In this regard, the impact of climate change has little significance on corrosion initiation and damage, though the mean carbonation depth is projected to increase. The projected mean carbonation depth reaches 19.8mm for A1FI emission scenario, 18.6 for A1B emission scenario, and 17.5mm for 550 stabilisation scenario by 2100, in comparison with 15.4mm estimated without the effect of climate change, but they are far less than the cover thickness to reach the reinforcement.

Bridge BC2 (1968, Hunter Region)

Bridge BC2 was constructed in 1968. The reinforced concrete at test site has 53mm cover with carbonation measured in the range of 5-25mm. The carbonation depth is still far less than the level that can induce corrosion of reinforcement. Climate change may affect the carbonation depth, but is unlikely to have an influence on the probability of corrosion initiation and damage. The carbonation depth is projected to be 32.2mm, 30.1mm and 28.3mm by 2100 for the three emission scenarios in comparison with 24.7mm in the absence of climate change.

3.2.4 1971-1994 Bridge Structures

Bridge BA3 was constructed in 1980 and Bridge BD3 was constructed in 1984. The concrete cover at test site is 58mm and 62mm, respectively. The measured carbonation depth is much smaller than the cover, implying that there is less significance in the occurrence of carbonation-induced corrosion.
4. **CLIMATE CHANGE IMPACT ON CORROSION OF EXISTING CONCRETE PORT STRUCTURES IN TROPICAL CLIMATE ZONES**

Concrete structures, including a slab soffit and a column in a berth of Port Townsville, Queensland, as shown in Figure 4-1, are considered for the case studies.

The berth was constructed circa 1960. It consists of a flat reinforced concrete deck approximately 530 mm thick, supported by circular reinforced concrete columns, as shown in Figure 4-2 (or the head photo of this chapter), generally 760 mm in diameter, with some 910 mm in diameter, on a 6.1 m square grid which are founded on pile caps which are in turn supported by driven universal piles. The berth and its adjacent have been constructed as a single rectangular wharf structure with the central portion of the concrete structure founded of fill. They appear to comprise the same reinforced concrete retaining walls, piles and slabs that extend around three sides of the wharf. The total wharf structure is approximately 250 m long and 70 m wide with the piled area for each berth is approximately 22m wide.

Following detailed inspections in 2008, the reinforced concrete of the berth is in a relatively good condition considering its age of 50 years (48 years in 2008). The results of the 2008 condition assessment indicated that with ongoing inspection and maintenance the deck should be serviceable for the next 50 years. The support columns for the berth were at risk of widespread deterioration within the next 10 years and are currently undergoing remedial works.
The test was done in 37 locations of the slab soffit with a cover of 50mm, and 8 locations of columns with a cover of 60mm. The test data of chloride concentration profiles and carbonation depth were provided by Port of Townsville, (2009). While chloride-induced corrosion is assessed on the location of slab soffits or columns with the highest chloride concentration, the carbonation-induced corrosion is assessed on the whole structures of the berth with different carbonation depths considered as an uncertainty in carbonation.
4.1 Chloride-Induced Corrosion of Port Structures

Slab Soffits

The tests were done for the concrete slab soffit between column 33 and 34 at column line N of the berth in Port Townsville in 2008. The slab has a cover of 50mm and is considered under exposure C2. The chloride content analysis shows the chloride concentration at 0.32%, 0.21%, 0.12% and 0.04 % by weight of concrete in the depth of 5mm, 20mm, 40mm and 50mm, respectively. With the concentration profile, the probability of chloride-induced corrosion initiation and damage is estimated at 10% and 9%, respectively, and likely mean rebar loss at 0.34mm.

Figure 4-3 Probability of chloride-induced corrosion initiation and damage and mean rebar loss of concrete structures of a slab soffit (33-34N) in a tropical climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘Year 2000 level’ is the relevant value in the absence of climate change.

As shown in Figure 4-3, the probability of corrosion initiation is projected to reach 59% by 2100 for A1FI emission scenario, 58% for A1B emission scenario and 57% for 550ppm stabilisation emission scenario in comparison with 55% in the absence of climate change. The probability of corrosion damage is 58%, 57% and 56% for the three emission scenarios respectively in comparison with 54% estimated without the effect of climate change. The mean rebar loss is project to be 9.3mm, 9mm and 8.8mm.
for three emission scenarios in comparison with 8mm without the effect of climate change.

In this regard, the impact of climate change by 2100 is limited within 4% increase in probability value of corrosion initiation and damage (or an increase of 7.4% in percentage), and it is within 1.3mm of mean rebar loss or a increase of 16%.

**Columns**

Column L5 of the berth in Port Townsville is selected as the case study of concrete port structures. It has cover of 60mm. The core sample from the column for chloride concentration analysis is located 500mm below tapered section of the column, and is considered at exposure C2. The chloride concentration analysis was done in 2008 and details the concentration profile that is 0.28%, 0.1%, 0.04%, 0% by the weight of concrete at the depth of 5mm, 20mm, 40mm and 60mm, respectively. Based on the chloride concentration profile, the probability of corrosion initiation and damage is less than 1% by 2008 and the mean rebar loss is estimated at 0.024mm.

As shown in Figure 4-4, the probability of corrosion of initiation and damage is projected less than 1.2% for all emission scenarios. The significance of climate change impact is very limited.

### 4.2 Carbonation-Induced Corrosion of Port Structures

**Slab Soffits**

The tests were done at 37 locations for the concrete slab soffit of the berth in Port Townsville in 2008. The slab has a cover of 50mm. The test of core sample from the slab shows carbonation depth in the range of 22mm to 55mm in 2008. On the basis of the test, it is very unlikely that there would be any occurrence of carbonation-induced corrosion initiation and damage by 2008.

As shown in Figure 4-5, the probability of corrosion initiation and damage as well as mean rebar loss may increase significantly over the years, and climate change may even magnify the change.
Figure 4-4 Probability of chloride-induced corrosion initiation and damage and mean rebar loss of concrete structures of column (L5) in a tropical climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘Year 2000 level’ is the relevant value in the absence of climate change.

As seen in the figure, by 2100, the probability of carbonation-induced corrosion initiation is projected to reach 91% for A1FI emission scenario, 89% for A1B emission scenario, and 86% for 550 ppm stabilisation emission scenario in comparison with 76% in the absence of climate change. In other words, climate change causes 15%, 13% and 10% increase in probability value. In the term of percentage change, the climate change leads to an increase of 20%, 17% and 13%, respectively. In fact, climate change leads to the largest impact on the probability of corrosion initiation around 2080, where the probability is 16% more than one estimated without the effect of climate change, or an increase of 31% in the term of percentage change.
Figure 4-5 Carbonation depth, probability of carbonation-induced corrosion initiation and damage and mean rebar loss of concrete structures of a column (L5) in a tropical climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘Year 2000 level’ is the relevant value in the absence of climate change.

Meanwhile, the probability of corrosion damage is 78%, 76% and 73% by 2100 for the three emission scenario, respectively, in comparison with 62% in the absence of climate change. In the term of percentage change, it is an increase of 26%, 23% and 18%, respectively. The mean rebar loss is 1.1mm, 1.0mm, 0.97mm for the three emission scenario in comparison with 0.79mm estimated without the effect of climate change.

Columns

Columns of the berth have a cover of 60mm. The core samples of the column were taken from 7 locations for carbonation depth assessment. The samples are located 500mm below tapered section of the column, showing a carbonation depth in the range of 22mm to 38mm by 2008. At such a depth of carbonation, it is very unlikely that the carbonation would cause any corrosion initiation and damage. As shown in Figure 4-6, without the effect of climate change, the carbonation depth is projected to reach 46.1mm by 2100. Although it is approaching the end of cover (60mm), it would still not cause significant change in the probability of corrosion initiation and damage.
As shown in the figure, the consideration of climate change significantly increases the probability of carbonation-induced corrosion, especially after 2060 for corrosion initiation and 2070 for corrosion damage. The probability of corrosion initiation is 45% for A1FI emission scenario, 28% for A1B emission scenario and 14% for 550 ppm stabilisation emission scenario, in comparison with only 1% in the absence of climate change. The probability of corrosion damage is projected to be 16%, 8.8% and 4.0% for the three emission scenarios in comparison with 0.17% estimated without the effect of climate change.

As a result, in this case, the climate change may have considerable impact on the reliability of concrete port structures. It is noted that the design of concrete cover is very critical to prevent from or reduce the impact of climate change.

Figure 4-6 Carbonation depth, probability of carbonation-induced corrosion initiation and damage and mean rebar loss of concrete structures of a column (L5) in a tropical climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘Year 2000 level’ is the relevant value in the absence of climate change.
5. ADAPTATION MEASURES TO COUNTERACT CLIMATE CHANGE IMPACT

Climate adaptation is considered as a process of deliberate change in anticipation of or in reaction to climatic stimuli and stress, in order to moderate or reduce adverse effects onto the system of our interest (IPCC, 2007). The critical component in climate adaptation is the adaptive capacity that measures the capability of the system to maintain its functionality and integrity under the disturbance of external stimuli and stresses. Enhancement of climate adaptive capacity, along with reducing vulnerability, is considered as one of prime approaches to counteract the changing climate.

For existing concrete structures under changing climate, the climate adaptation in terms of the enhancement of adaptive capacity can be done by developing new technologies for maintenance to counter the impact of increasing corrosion risk under changing climate. On the other hand, there is a wide range of conventional maintenance and retrofitting options that can enhance the durability of concrete structures and these can be applied to reduce the adverse affects of climate change. The maintenance options generally include surface coating, realkalisation, chloride extraction, cathodic protection, and cover replacement.

Creating a surface barrier by coating is more appropriate for reducing the exposure of concrete structure to external stimuli. Meanwhile, extraction and cathodic protection is more commonly carried out for structures with high corrosion risk to reduce the ingress of deleterious agents. The cover replacement is most effective, but also the most expensive options, followed by cathodic protection that also has operating cost and then
realkanisation or chloride extraction. The surface coating is the cheapest option, but is also less effective.

5.1 Simulation of Adaptation Measures for Chloride- and Carbonation-Induced Corrosion

To consider the worst scenario of climate change impact, A1FI emission scenario is applied to assess the adaptation of concrete structures for chloride-induced corrosion. Meanwhile, as indicated previously, the simulation of adaptation measures for chloride- and carbonation-induced corrosion is represented by various factors.

It is assumed in the simulation of adaptation measures for chloride-induced corrosion:

1. Electrochemical Chloride Extraction 90%: correction factor for chloride concentration $R_{\text{Chloride}} = 0.1$, and correction factor for corrosion rate $R_{\text{corr}} = 0.1$ accordingly. The cost is about $600/m^2$. It is a permanent solution if a surface coating is applied and properly maintained.

2. Polyurethane sealer: It is able to reduce chloride diffusion coefficient. The estimation of diffusion coefficient is described in section 2.4.2, and relevant data described in Table 2-2. The cost is about $40/m^2$. The coating is applied every 15 years.

3. Polymer-modified (p-m) cementious coating: it is able to reduce chloride diffusion coefficient. The estimation of diffusion coefficient is described in section 2.4.2, and relevant data described in Table 2-2. The cost is about $40/m^2$. The coating is applied every 15 years.

4. Replacing existing cover: the new cover conforms to current concrete standards, as described in section 2.4.5. The cost is about $2,500/m^2$.

5. Cathodic protection: it permanently stops corrosion initiation and damage. The cost includes $800/m^2$ initial cost and $10/m^2/\text{year}$ operating cost.

At the same time, it is assumed in the simulation of adaptation measure for carbonation-induced corrosion:

6. Realkalisation: It is a permanent solution if a surface coating is applied and properly maintained. The cost is about $600/m^2$.

7. Acrylic-based coating: It is able to reduce carbonation diffusion coefficient. The estimation of diffusion coefficient is described in section 2.4.2, and relevant data described in Table 2-2. The cost is about $50/m^2$. The coating is applied every 15 years.

8. Replacing existing cover: the new cover conforms to current concrete standards, as described in section 2.4.5. The cost is about $2,500/m^2$.

Adaptation measures are assumed to be applied in 2011 in the simulation, and all cost is the present value in 2011.
5.2 Cost/Benefit Assessment of Adaptation Options

For adaptation measures, the residual risk is defined by:

\[
\text{Risk}(T) = \sum_{i=1}^{T/\Delta t} p_s(i\Delta t) - p_s([i-1]\Delta t) \quad \frac{1 - p_s([i-1]\Delta t)}{1 - p_s(i\Delta t)}
\]  

where \(p_s(i\Delta t)\) is the cumulative probability of corrosion damage at time \(i\Delta t\) (\(i=1,T/\Delta t\)), and \(T\) is the service life. Since there is lack of information of damage cost, a proxy of the benefit due to the implementation of adaptation options in changing climate is defined by:

\[
\text{Benefit} = \text{Risk}(T, \text{BAU} | \text{CC}) - \text{Risk}(T, \text{Adaptation} | \text{CC})
\]  

where \(\text{Risk}(T, \text{BAU} | \text{CC})\) is the risk in the circumstance of ‘business as usual’ or ‘do-nothing’ under climate change, and \(\text{Risk}(T, \text{Adaptation} | \text{CC})\) is the residual risk after the implementation of an adaptation option under climate change. CC represents the effect of climate change, and NCC represents no climate change considered.

Eqn.(5-2) can be rewritten as:

\[
\text{Benefit} = [\text{Risk}(T, \text{BAU} | \text{CC}) - \text{Risk}(T, \text{BAU} | \text{NCC})] + [\text{Risk}(T, \text{Adaptation} | \text{NCC}) - \text{Risk}(T, \text{Adaptation} | \text{CC})]
\]  

In Eqn. (5-3), the first term represents the benefit contributed by an adaptation option to mitigate the risk of corrosion completely due to climate change, the second term gives the benefit to increase corrosion resistance or adaptive capacity by the maintenance in the absence of climate change (the maintenance option is the same as adaptation option, applied to mitigate corrosion risk as usual but nothing to do with climate change), and the third term describes the loss of the benefit due to the reduced effectiveness of the adaptation option as a result of climate change. In another word, the cost is to achieve the benefit and to offset the loss of effectiveness of adaptation options due to climate change.

Meanwhile, the adaptation cost is estimated on the basis of initial and operating cost to maintain the performance or function of adaptation measures given a discount rate.

In addition, ‘Benefit’ as described in Eqn.(5-3) also represents the degree of the reduction of corrosion risk when an adaptation option is applied, and can thus be considered as the representation of adaptation effectiveness. An option is not recommend when ‘Benefit\(\leq 0\)’. The maximum benefit or adaptation effectiveness can be achieved when the risk, or \(\text{Risk}(T, \text{Adaptation} | \text{CC})\) in Eqn.(5-1), approaches zero after an adaptation option implemented.
Meanwhile, based on the cost and the effectiveness, an adaptation effectiveness diagram, such as shown in Figure 5-1, can be established to inform decision-makers to select proper adaptation options based on both effectiveness and cost. In the diagram, each bar represents an adaptation option with its width representing adaptation effectiveness and height representing cost. The optimal option is low cost and great effectiveness.

![Adaptation Effectiveness Diagram](image)

Figure 5-1 Conceptual illustration of adaptation effectiveness diagram

5.3 Adaptation Assessment for Chloride-Induced Corrosion and Cost/Benefit

5.3.1 Bridge BB1 Constructed in 1925 in Sydney

Figure 5-2 describes the effect of adaptation measures (1)-(4) described above on the probability of corrosion initiation and damage as well as rebar loss. It appears the most effective approach is the replacement of old 29mm cover of concrete with new concrete that meets the current design stands of 65mm cover at exposure C2. The measure not only counteract the impact of climate change, but also enough to maintain the probability of corrosion initiation and damage at a very low level and almost no further rebar loss.
Corrosion initiation probability

Figure 5-2 Probability of chloride-induced corrosion initiation and damage and mean rebar loss of concrete structures of Bridge BB1 (1925) in NSW considering adaptation options (1)-(4). ‘Impact’ - with the effect of climate change; ‘Year 2000 level’ – the relevant value in the absence of climate change.

The second most effective approach is chloride extraction that assumes that 90% of chloride was removed. Although it is still projected that there is an increase in the probability of corrosion initiation and damage, but it maintain the probability of corrosion initiation below the estimation in the absence of climate change until 2080, and maintain the probability of corrosion damage below the projection in the absence of climate change. In another word, it has at least mitigated the impact of climate change.

The use of p-m cementious coating may also maintain the probability of corrosion initiation and damage below the estimation in the absence of climate change, while relatively less effective. In addition, polyurethane sealer, as indicated in Table 2-2, has little effect on the probability and mean rebar loss, and is thus not recommended for use.

Considering a service life of concrete structure till 2100, the cost and cost/benefit ratio is shown in Figure 5-3. It should be pointed out that the cost/benefit ratio is not in the sense of the ratio of dollar to dollar, and it is just a relative indicator for the comparison. As shown in the figure, cover replacement is most expansive and also very high in cost/benefit ratio among the five options, while p-m cementious coating is the least. The
cost/benefit ratio of the implementation of polyurethane sealer is most sensitive to the discount rate.

![Figure 5-3: Cost and cost/benefit in relation to discount rate for the implementation of adaptation options for chloride-induced corrosion of concrete structures of Endeavour Bridge (1925) in NSW under climate change.](image)

The cost is contributed to reduce corrosion damage risks as usual and the increased risks due to climate change as well. On the basis of Eqn.(5-3), the cost contributes to three part, 1) mitigating corrosion risk due to climate change, 2) increasing corrosion resistance, and 3) offsetting the loss of adaptation effectiveness due to climate change. Table 5-1 gives the subdivision of the cost related to the three parts.

<table>
<thead>
<tr>
<th>Adaptation/Maintenance Measures</th>
<th>Percentage of total cost to mitigate increased risk due to climate change</th>
<th>Percentage of the total cost to increase corrosion resistance</th>
<th>Percentage of the total cost to offset the loss of adaptation effectiveness due to climate change</th>
</tr>
</thead>
<tbody>
<tr>
<td>Polyurethane sealer</td>
<td>47%</td>
<td>7%</td>
<td>46%</td>
</tr>
<tr>
<td>p-m cementious coating</td>
<td>31%</td>
<td>46%</td>
<td>23%</td>
</tr>
<tr>
<td>Chloride extraction</td>
<td>29%</td>
<td>53%</td>
<td>18%</td>
</tr>
<tr>
<td>Replacement</td>
<td>22%</td>
<td>78%</td>
<td>0%</td>
</tr>
<tr>
<td>Cathodic protection</td>
<td>22%</td>
<td>78%</td>
<td>0%</td>
</tr>
</tbody>
</table>

For an example, for p-m cementious coating, 31% of the total cost is contributed to the mitigation of increased corrosion risk due to climate change, 46% is contributed to improve the resistance of concrete structures to chloride attack, and 23% is contributed to the loss of the effectiveness of adaptation options due to climate change. For polyurethane, nearly a half of cost is actually for offsetting the loss of its effectiveness due to climate change.
It should be pointed out that the cathodic protection in the table is assumedly not affected by the climate change. In fact, its effectiveness may be affected by sea level rise, which will not be discussed in this report.

Finally, Figure 5-4 indicates the overall effectiveness of adaptation options under changing climate in relation to the cost considering a discount rate of 0.03. The width of the bar represents the relative adaptation effectiveness of each option. Polyurethane sealer is least effective.

Figure 5-4 Cost and effectiveness of adaptation options for chloride-induced corrosion of concrete structures of Bridge BB1 (1925) in NSW, at a discount rate of 0.03.

5.3.2 Bridges BD2 Constructed in 1967 in Northern Region

Figure 5-5 describes the effect of adaptation measures (1)-(4) described above on the probability of corrosion initiation and damage as well as rebar loss. It appears the most effective approach is the use of p-m cementious coatings on the surface of 62mm cover of concrete that already meets the current design stands for structures at exposure C2. The measure not only counteracts the impact of climate change, but also enough to maintain the probability of corrosion initiation and damage at a reasonably low level.
Figure 5-5 Probability of chloride-induced corrosion initiation and damage and mean rebar loss of concrete structures of Bridge BD2 (1967) in northern region NSW considering adaptation options (1)-(4). ‘Impact’ - with the effect of climate change; ‘Year 2000 level’ – the relevant value in the absence of climate change.

The second most effective approach is cover replacement that assumes the same cover thickness but at lower rate of penetration. Although it is still projected that there is an increase in the probability of corrosion initiation and damage, but it maintains the probability of corrosion initiation below the estimation in the absence of climate change until 2080, and maintains the probability of corrosion damage below the projection in the absence of climate change. In others word, it has at least mitigated the impact of climate change.

Considering a service life of concrete structure till 2100, the cost and cost/benefit ratio is shown in Figure 5-3. Once again, the ratio is just a relative indicator for the purpose of comparison. As shown in the figure, cover replacement is most expensive and also highest cost/benefit among the five options, while p-m cementious coating is the least. Different from the last case, the use of polyurethane coating appears also much lower in addition to its lower sensitivity to the discount rate.
Figure 5-6 Cost and cost/benefit in relation to discount rate for the implementation of adaptation options for chloride-induced corrosion of concrete structures of Bridge BD2 (1967) in northern regions of NSW under climate change.

Table 5-2 gives the subdivision of the cost related to the three parts. Comparing to the previous case, polyurethane sealer is much more effective with 29% of the total cost is contributed to the mitigation of increased corrosion risk due to climate change, 57% is contributed to improve the resistance of concrete structures to chloride attack, and moderate 14% is contributed to the loss of the effectiveness of adaptation options due to climate change. In general, more than a half of the total cost goes to increase the resistance to corrosion.

Table 5-2 Subdivision of the cost of the adaptation/maintenance measures for chloride-induced corrosion of Bridges BD2 (1967) in the northern region of NSW.

<table>
<thead>
<tr>
<th>Adaptation/Maintenance Measures</th>
<th>Percentage of total cost to mitigate increased risk due to climate change</th>
<th>Percentage of the total cost to increase corrosion resistance</th>
<th>Percentage of the total cost to offset the loss of adaptation effectiveness due to climate change</th>
</tr>
</thead>
<tbody>
<tr>
<td>Polyurethane sealer</td>
<td>29%</td>
<td>57%</td>
<td>14%</td>
</tr>
<tr>
<td>p-m cementious coating</td>
<td>22%</td>
<td>74%</td>
<td>4%</td>
</tr>
<tr>
<td>Chloride extraction</td>
<td>26%</td>
<td>64%</td>
<td>10%</td>
</tr>
<tr>
<td>Replacement</td>
<td>25%</td>
<td>69%</td>
<td>6%</td>
</tr>
<tr>
<td>Cathodic protection</td>
<td>22%</td>
<td>78%</td>
<td>0%</td>
</tr>
</tbody>
</table>

Once again, it should be pointed out that the cathodic protection in the table is assumedly not affected by the climate change. However, its effectiveness may be affected by sea level rise, which will not be discussed in this report.

In addition, Figure 5-7 indicates the overall effectiveness of adaptation options under changing climate in relation to the cost with considering a discount rate of 0.03.
figure, the width of each bar represents the relative adaptation effectiveness of each option, while the height represents the cost. For this case, p-m cementious coating becomes a very attractive option with low cost and high effectiveness. On the other hand, cover replacement is least attractive with high cost and relatively low effectiveness.

Figure 5-7 Cost and effectiveness of adaptation options for chloride-induced corrosion of concrete structures of Bridge BD2(1967) in the northern region of NSW, at a discount rate of 0.03.

5.3.3 Concrete Slab Soffits in Port Townsville

Figure 5-8 describes the effect of adaptation measures (1)-(4) described above on the probability of corrosion initiation and damage as well as rebar loss of concrete slabs in one of berths in Port of Townsville. It appears the most effective approach to reduce the corrosion is the use of either cover replacement or p-m cementious coatings, though the probability corrosion initiation and damage increase again to almost 30% by 2100 under A1FI emission scenario.

Other approaches including chloride extraction that only marginally reduces the probability of corrosion initiation and damage. The use of polyurethane sealer cannot even maintain the probability below the estimation in the absence of climate change.
4B ADAPTATION MEASURES TO COUNTERACT CLIMATE CHANGE IMPACT

Figure 5-8 Probability of chloride-induced corrosion initiation and damage and mean rebar loss of structures of concrete slab in Port of Townsville considering adaptation options (1)-(4). ‘Impact’ - with the effect of climate change; ‘Year 2000 level’ – the relevant value in the absence of climate change.

Considering a service life of concrete slab structure till 2100, the cost and cost/benefit ratio is shown in Figure 5-9. As seen, cover replacement is most expansive and also highest cost/benefit among the five options, while p-m cementious coating is the least. It also indicates that the cost/benefit ratio to use polyurethane coating appears sensitive to the discount rate.

Table 5-3 gives the subdivision of the cost related to the three parts. As seen, polyurethane sealer show less effective with 36% of the total cost contributed to the loss of the effectiveness of adaptation options due to climate change. Meanwhile, 38% of the total cost is contributed to the mitigation of increased corrosion risk due to climate change, and 25% is contributed to improve the resistance of concrete structures to chloride attack. In general, except of polyurethane sealer, more than a half of the total cost of all options goes to increase the resistance to corrosion, especially cathodic protection and p-m cementious coating that is as high as 87% and 75%, respectively. In another word, the adaptive capacity will be significantly strengthened to resist chloride-induced corrosion.
Figure 5-9 Cost and cost/benefit in relation to discount rate for the implementation of adaptation options for chloride-induced corrosion of a concrete slab in Port of Townsville under climate change.

Table 5-3 Subdivision of the cost of the adaptation/maintenance measures for chloride-induced corrosion of concrete slab of port structure in the port of Townsville.

<table>
<thead>
<tr>
<th>Adaptation/Maintenance Measures</th>
<th>Percentage of total cost to mitigate increased risk due to climate change</th>
<th>Percentage of the total cost to increase corrosion resistance</th>
<th>Percentage of the total cost to offset the loss of adaptation effectiveness due to climate change</th>
</tr>
</thead>
<tbody>
<tr>
<td>Polyurethane sealer</td>
<td>38%</td>
<td>25%</td>
<td>36%</td>
</tr>
<tr>
<td>p-m cementious coating</td>
<td>16%</td>
<td>75%</td>
<td>9%</td>
</tr>
<tr>
<td>Chloride extraction</td>
<td>28%</td>
<td>51%</td>
<td>21%</td>
</tr>
<tr>
<td>Replacement</td>
<td>20%</td>
<td>71%</td>
<td>8%</td>
</tr>
<tr>
<td>Cathodic protection</td>
<td>13%</td>
<td>87%</td>
<td>0%</td>
</tr>
</tbody>
</table>

As discussed before, the cathodic protection in the table is assumedly not affected by the climate change, which may indeed be affected by sea level rise. However, it will not be discussed in this report.

Figure 5-10 describes the overall effectiveness of adaptation options under changing climate in relation to the cost with considering a discount rate of 0.03. In the figure, the width of each bar represents the relative effectiveness of each adaptation option, while the height represents the cost. For this case, cathodic protection and p-m cementious coating becomes two options to effectively mitigate chloride-induced corrosion under changing climate.
5.4 Adaptation Assessment for Carbonation-Induced Corrosion and Cost/Benefit

5.4.1 Concrete Slab Soffits in Port Townsville

Figure 5-11 describes the effect of adaptation measures (6)-(8) introduced above on the probability of corrosion damage of concrete slabs in one of berths in Port of Townsville. Considering that the acrylic-based surface coating is normally applied in dry environment, only two options, realkanisation and cover replacement, are considered for the concrete slab at exposure C2. Realkalisation is an electrochemical process to raise pH near reinforcement. After realkalisation, concrete will not easily recarbonate. It is therefore assumed that corrosion will stop permanently. Meanwhile, cover replacement also stops corrosion while it may not stop carbonation.

Considering a service life of concrete slab structure till 2100, the cost and cost/benefit ratio is shown in Figure 5-9. As seen, cover replacement is most expansive and also highest cost/benefit among the five options, while p-m cementious coating is the least. It also indicates that the cost/benefit ratio to use polyurethane coating appears sensitive to the discount rate.
Figure 5-11 Probability of carbonation-induced corrosion damage of structures of concrete slab in Port of Townsville considering adaptation options (6) and (8). ‘Impact’ - with the effect of climate change; ‘Year 2000 level’ – the relevant value in the absence of climate change.

(g) Cost

(h) Cost/Benefit

Figure 5-12 Cost and cost/benefit in relation to discount rate for the implementation of adaptation options for carbonation-induced corrosion of a concrete slab in Port of Townsville under climate change.

Table 5-4 gives the subdivision of the cost related to the three parts. As seen, both alkanisation and cover replacement 38% of the total cost contributed to the mitigation of increased corrosion risk due to climate change, and other 62% is contributed to improve the resistance of concrete structures to chloride attack or adaptive capacity. However, their cost to reach that effectiveness is different.

Figure 5-13 describes the overall effectiveness of adaptation options under changing climate in relation to the cost with considering a discount rate of 0.03. In the figure, the width of each bar represents the relative effectiveness of each adaptation option for
carbonation-induced corrosion, while the height represents the cost. For this case, realkanisation has the advantage over the cover replacement to effectively mitigate chloride-induced corrosion under changing climate due to its low cost.

Table 5-4 Subdivision of the cost of the adaptation/maintenance measures for carbonation-induced corrosion of concrete slab of port structure in the port of Townsville.

<table>
<thead>
<tr>
<th>Adaptation/Maintenance Measures</th>
<th>Percentage of total cost to mitigate increased risk due to climate change</th>
<th>Percentage of the total cost to increase corrosion resistance</th>
<th>Percentage of the total cost to offset the loss of adaptation effectiveness due to climate change</th>
</tr>
</thead>
<tbody>
<tr>
<td>Realkanisation</td>
<td>38%</td>
<td>62%</td>
<td>0%</td>
</tr>
<tr>
<td>Cover replacement</td>
<td>38%</td>
<td>62%</td>
<td>0%</td>
</tr>
</tbody>
</table>

Figure 5-13 Cost and effectiveness of adaptation options for carbonation-induced corrosion of concrete slab structures in the Port of Townsville at a discount rate of 0.03.
6. SUMMARY

Since most designs of existing concrete structures do not take into account the effect of a changing environment, there are concerns that many existing concrete infrastructure are likely to suffer from decreased durability, safety and serviceability.

This report has established a probabilistic simulation approach based on chloride- and carbonation-induced corrosion models that can be calibrated by the historical field measurement of chloride concentration and carbonation penetration and then can project the probability of corrosion initiation and damage into future. We developed models of adaptation options and looked at their effects on various parameters of corrosion models for existing concrete structures, such as carbonation depth and carbon dioxide diffusion coefficient for carbonation-induced corrosion, and chloride concentration, chloride diffusion coefficient, critical chloride concentration for chloride-induced corrosion, and corrosion rate. These models were then applied to simulate the corrosion under the influence of adaptation measure, including surface coating, alkanisation, chloride extraction, cover replacement, cathodic protection and so on. The effect of climate change is also considered in the models through the influence of carbon dioxide concentration, temperature and relative humidity.

Two typical concrete structures, bridges and port structures, are used as case studies to elucidate the impact of climate change on existing concrete structures and available options as well as their effectiveness in reducing the impact of climate change and increasing adaptive capacity to counteract corrosion. The investigated bridges are located in Sydney, Southern, Northern and Hunter regions of NSW and constructed during the period of pre-1959, 1959-1970, 1971-1994 and post-1995. At the same time, the investigated port structures are concrete slab soffits and columns from a berth that is managed by Port of Townsville Limited. As found, although all investigated structures
are under exposure C2, the concrete cover does not necessarily meet the current standard, such as AS 3600-2009, that requires 65mm. In fact, the cover of the most of existing bridges in this case study is less than 65mm, with one reaching as low as 29mm. This explained more on the importance to specifically assess durability of existing bridges and especially when considering the impact of climate change on the durability in relation to carbonation- and chloride-induced corrosion.

In the part 2 of the report, it is indicated that the effect of climate change on chloride-induced corrosion of concrete structures which design follows the AS3600 and AS5100.5, is within an increase of 3.5% in probability value by 2100. In practice, the change of the probability can go higher due to non-binding on the standards or lack of quality assurance in construction, for example, use of lower concrete cover. The bridges in Sydney, constructed in 1925 with 29mm cover, shows that the climate change at A1FI emission scenario can leads up to 8% increase in corrosion initiation and damage probability value, or 13% increase in percentage change. Even for a modern bridge in the northern region of NSW, constructed in 1984, the structure on the bridge at exposure C1 and C2 may experience up to 5-7% increase in probability value by 2100 in the presence of climate change impact at A1FI emission scenario.

Due to an improper concrete cover of some of the early constructed bridges, climate change may lead to a considerable impact on carbonation-induced corrosion. For example, the bridge constructed in 1925 in Sydney has only 29mm concrete cover. The probability of corrosion initiation is 72%, 67% and 63% by 2100 for A1FI, A1B and 550 ppm stabilisation emission scenarios, respectively, in comparison with 51% estimated in the absence of climate change. In another word, the probability increases 21%, 16% and 12% in value, or an increase of 41%, 31% and 14% in percentage change due to climate change. Meanwhile, the probability of corrosion damage is 63%, 60% and 55% in comparison with 44% estimated in the absence of climate change, also a significant increase due to climate change. For a concrete column at exposure C2 in Port Townsville, climate change impact may lead to an increase of corrosion initiation probability form 1% to 45% for A1FI emission scenario, 28% for A1B emission scenario and 14% for 550 ppm stabilisation emission scenario by 2100. It also leads to the increase corrosion damage probability from 0.2% to 16%, 8.8% and 4.0% for the three emission scenarios by 2100.

Both chloride-induced and carbonation-induced corrosion show the potential experience of a scalable impact of climate change, which should be considered for maintenance planning. Adaptation options should also be developed and optimised to mitigate the impact and enhance the adaptive capacity of concrete structures to changing climate.

In the simulation of implementation of adaptations to counteract the impact of climate change, five options including electrochemical chloride extraction, polyurethane sealer, polymer-modified cementious coating, cover replacement and cathodic protection, were considered to reduce chloride-induced corrosion. Other two options including realkalisation and cover replacement were introduced to mitigate carbonation-induced corrosion. Meanwhile, cost and adaptation effectiveness are also introduced to quantify the adaptation options for the identification of the most preferable option in association with a specific concrete structure, such as a slab or column. The cost includes initial
implementation cost and on-running operating cost, which are all converted to their present value in 2011 with a discount rate ranging from 0.01 to 0.1 selected for sensitivity assessment. The effectiveness, also known as a proxy of benefit due to the implementation of adaptation options, is defined as the amount of reduction in corrosion risk from ‘business as usual (BAU)’ after implementing adaptations. As a result, adaptation effectiveness diagram is developed by combining cost with adaptation effectiveness, which may facilitate the decision-making in developing adaptation strategies that maximise adaptation effectiveness with minimised cost.

The cost contributes to three factors, i.e. 1) reducing the impact of climate change, 2) increasing adaptive capacity to resist corrosion, and 3) offsetting the loss of adaptation effectiveness due to climate change. In general, the more there is more on offsetting, less effective is the option. The case study of concrete bridges indicates that the replacement of concrete cover is often to be the most effective option, but it is also the most expensive one. Surface coating is the least costly, but is usually, but not always relatively less effective. For the bridge constructed in 1925 in Sydney, cathodic protection is the preferred adaptation measure to mitigate the chloride-induced corrosion damage due to its greater effectiveness and moderate cost. Among the total cost, 22% is contributed to mitigate the increase corrosion damage risk due to climate change, 78% is contributed to increase adaptive capacity to resist corrosion, and nothing is contributed to offset the loss of adaptation effectiveness due to climate change. It should be pointed out the effectiveness of cathodic protection is in fact affected by sea level rise that may change the cost for offset the loss of effectiveness. At the same time, the use of polyurethane sealer is the least preferred due to its very low effectiveness though low cost. For this, 46% of the total cost is contributed to offset the loss, which is not really beneficial to the enhancement of adaptive capacity to counteract corrosion damage.

Depending on residual risk of corrosion damage of concrete structure after implementing adaptation options, the preferred adaptation option can vary. For a bridge constructed in 1967 in the northern region of NSW, polymer-modified cementious coating appears the most preferable due to its great effectiveness with 24% of the total cost is contributed to mitigate the increasing risk as a result of climate change, 57% is contributed to strengthen the adaptive capacity, and only 4% is contributed to offset the loss of adaptation effectiveness, in comparison with 14% for polyurethane sealer, 10% for cathodic protection. Meanwhile, it has a cost much lower than cover replacement, cathodic protection and chloride extraction.

A similar approach was applied for the cost/benefit assessment of adaptation options for carbonation-induced corrosion of port structures in relation to realkalisation and cover replacement are considered.

As reported in part 2 of the report, the climate change impact assessment on the aspect of design that follows the Australian standards may provide general rules for concrete structural design taking into account effects of changing climate. Different from that, impacts on existing concrete infrastructure and the adaptation that should be applied to mitigate the impact are specific due to the uniqueness of individual structures especially regarding their different local environment exposure history as well as the uncertainties
in construction and maintenance. Therefore, an effective adaptation option should be
developed at the level of individual concrete structures. Finally, cost and benefit
assessment should further be developed to consider the lifecycle of concrete
infrastructure.
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Part 3: Case Studies of Concrete Deterioration and Adaptation (Brief)
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Part 2: Modelling and Simulation of Deterioration and Adaptation Options;
Part 3: Case Studies of Concrete Deterioration and Adaptation.

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1. INTRODUCTION

The residual service life of existing concrete structures is largely determined by its deterioration over time. The deterioration rate of concrete structures depends not only on the construction processes employed and the composition of the materials used in the construction process, but also on the current as well as past environment. Meanwhile, climate change may alter this environment in the future, especially in the long term, causing more acceleration of deterioration processes and consequently affecting the safety and serviceability of existing concrete infrastructure. In particular, many existing concrete structures, for which the design has not taken into account the effect of changing climate, are likely to suffer from more decreased durability as a consequence of climate change and incur more damage and maintenance cost. Considering the amount of existing concrete infrastructure, the potential impacts of the climate change cannot be ignored.

As indicated in the previous part report, the deterioration of concrete can be affected directly or indirectly by climate change impacts, in association with the change in carbon dioxide (CO₂) concentration, temperature and relative humidity, as shown in Table 1-1. The climate-related deterioration of concrete structures is mostly caused by the infiltration of deleterious substances from the environment, for example carbon dioxide and chloride, which causes reinforcement corrosion.

Understanding the implications of climate change on existing concrete structure is vital for effective decision-making in asset management to protect concrete buildings and infrastructure that underpin human settlements and the economy. In practice, the
durability and serviceability of concrete structures are maintained via routine inspection, maintenance and replacement. Climate change impacts needs to be considered in both the application and effectiveness of maintenance and replacement regimes. Considering the extensive uncertainties and the limited knowledge of future climate, simulation is once again deemed as an effective approach that may provide insights into how likely and how much the future climate would impact on existing concrete structures. More importantly, taking the precautionary principle, simulations may inform the necessary extent of change in maintenance required to maintain the safety, serviceability and durability at given the likely climate change scenarios.

<table>
<thead>
<tr>
<th>Climate Change</th>
<th>Implications</th>
</tr>
</thead>
<tbody>
<tr>
<td>Increase of carbon concentration</td>
<td>Elevated carbon concentration accelerates carbonation and increases carbonation depth in concrete: this increases the likelihood of concrete structures exposed to carbonation induced reinforcement corrosion initiation and structural damage</td>
</tr>
<tr>
<td>Change of temperature</td>
<td>Elevated temperature accelerates carbonation, chloride penetration and corrosion rate of reinforcement that exacerbates the corrosion initiation and structural damage</td>
</tr>
<tr>
<td>Change of humidity</td>
<td>Lowered relative humidity may reduce or even stop carbonation and chloride penetration in the area with yearly average RH currently just above 40-50%, while increased humidity may result in them occurring in the regions where they are now negligible.</td>
</tr>
</tbody>
</table>

This report will focus on the modelling and simulations for existing concrete infrastructure, and provide an effective way to explore adaptation options to reduce or mitigate the climate change impacts by case studies of concrete bridges and port structures.

In Chapter 2, we will establish methodology and present modelling and simulation for existing concrete deterioration based on field testing data integrated with probabilistic approaches. Models for the simulation of various adaptation measures for chloride and carbonation induced corrosion will also be introduced in the chapter. Using modelling and simulation, cases studies will be carried out in Chapter 3 in relation to the climate change impact on existing concrete bridges constructed in the period of pre-1959, 19959-1970, 1971-1994 in Sydney, Southern, Northern as well as Hunter regions in NSW that are located within a temperate climate zone in Australia. The impact on chloride-induced corrosion will be represented by changes in mean corrosion initiation probability, corrosion damage probability and mean rebar loss. Climate change impact on carbonation-induced corrosion will also be addressed through the same concrete
structures in NSW. In Chapter 4, a similar assessment will be carried out for concrete slabs and columns of port structures in the Port of Townsville that is located within a tropical climate zone in Australia. In Chapter 5, adaptation options to counteract or mitigate climate change impact on both chloride and carbonation induced corrosion will be discussed, and their cost and benefit will be presented in relation to the mitigation of climate change impact, increase of adaptive capacity and offset of adaptation effectiveness loss. Meanwhile, adaptation effectiveness diagrams will be developed for each case for decision-making on the selection of adaptation options in order to most cost-effectively maintain the durability and serviceability of bridges and port structures, which can be further extended to assess other concrete infrastructure.
2. SIMULATION AND MODELLING OF EXISTING CONCRETE STRUCTURES

Simulation and modelling of carbonation and chloride induced corrosion have been discussed in great detail in the second part of the report. These simulations and models are based on the assumption that the design of concrete structures meets the Australian standard of AS3600 – Concrete Structures. Although these simulations and models are able to effectively project the overall performance of general deteriorating concrete structures from the aspect of a new design with consideration of changing climate, they could be inaccurate for the assessment of the performance of an individual existing concrete structure that can be different in practice as a result of uncertainties in construction and local environment as well.

The simulation of carbonation and chloride penetration induced corrosion of existing concrete is implemented by a conventional probabilistic approach on the basis of Monte-Carlo simulation, as demonstrated in the flow chart described in Figure 2-1. This approach considers environmental variables and their uncertainties, including concentration of carbon dioxide, yearly mean temperature and relative humidity. Different from the simulation described in the part 2 report, modelling of deterioration of existing individual concrete structures under changing climate should be calibrated by field testing data, as illustrated in the flow chart, so that corresponding climate adaptations can be properly developed. In this chapter, the procedure to modify the models developed in the previous report will be introduced to consider the characteristics of specific concrete structures. Meanwhile, the modelling of adaptations will also be developed for the assessment of their effectiveness to reduce the impact of climate change.
2.1 Carbonation-Induced Corrosion Modelling with Calibration by Field Testing Data

As discussed in the Part 2 – Modelling and Simulation of Deterioration and Adaptations, carbonation depth at calendar year $t$ is estimated based on theory modified from (Yoon et al 2007) in relation to mass concentration of ambient $CO_2$, diffusion coefficient in concrete, the age factor for the $CO_2$ diffusion coefficient, cement content, $CaO$ content in cement, degree of hydration, the age factor for microclimatic conditions associated with the frequency of wetting and drying cycle, the $CO_2$ urbanisation factor, and temperature factor. For an existing concrete structure, the parameters related to the theory are calibrated by carbonation depth measurements.

As known, carbonation is highest for relative humidity (RH) = 40-75%, or 50-70% (Russell et al 2001). Al-Khaiat and Fattuhi (2002) report that little or no carbonation occurs below a RH of 30%, whereas Russell et al (2001) state that below 50% RH there is insufficient moisture for carbonation reactions to take place. Most carbonation models assume RH>50%. To be conservative, assume that if RH($t$) < 40% then...
carbonation front ceases to advance, or carbonation depth does not increase. Relative humidity prior to 2000 is likely to be well over 40% for most infrastructure locations in Australia, otherwise there would be no carbonation and no corrosion - and so no test results to analyse. Hence, relative humidity is not considered in the development of diffusion coefficient from test data.

Carbonation depths obtained using a phenolphthalein pH indicator are measured to an accuracy of ±0.5 mm (e.g., Al-Khayat et al. 2002, Jones et al. 2000). It is recognised that carbonation test methods may result in variability in test measurements (e.g., Jones et al. 2000). However, test measurements based on averaging several sample results should minimise these uncertainties. The present analysis assumes that the carbonation depth obtained using a phenolphthalein pH indicator is an accurate indicator of carbonation.

It is observed that corrosion may occur when the distance between the carbonation front and the reinforcement bar surface is less than 1-5 mm (e.g., Yoon et al 2007). However, probabilistic analyses for assessing durability design specifications tend to ignore this effect (Duracrete 2000b, fib 2006). Hence, time to corrosion initiation occurs when carbonation front equals concrete cover, and the probability of corrosion initiation is described by the likelihood of carbonation front equal to concrete cover.

2.2 Chloride-Induced Corrosion Modelling with Calibration by Field Testing Data

There is evidence to suggest that chloride action is accelerated by carbonation (and SO₂, NOₓ) because carbonation disturbs the equilibrium between free and bound chlorides in the concrete, thereby increasing the free chloride concentration in the pore solution. However, it appears that this evidence has not been translated into any useful quantitative models. Thus the interaction effect between carbonation and chlorides is, for the time being, omitted from the present study. Use total (acid soluble) chloride content.

Chloride penetration is described based on diffusion equation or Fick’s law, which is related to surface chloride concentration, initial chloride concentration that already exists in the concrete mix, chloride diffusion coefficient. Non-linear least squares methods are used to determine the ‘best fit’ of Fick’s law parameters for diffusion coefficient and surface chloride concentration with measurements from specimens. The specimens or cores is typically of 10 to 20 mm depth, and the mid-height value of each slice is used when determining chloride profiles at time t (e.g., McGee 1999, Tamimi et al 2008).

2.3 Modelling of Adaptations for Existing Structures

An adaptation may influence carbonation depth, chloride diffusion coefficient, chloride concentration, corrosion rate, or critical chloride concentration if the existing concrete
cover is not replaced. If the existing concrete cover is replaced with new concrete then in this case the deterioration process of the existing structure ceases and the deterioration process restarts at the time of cover replacement at time $t_{\text{adapt}}$.

It assumes that any coatings are adequately maintained during the life of the structure (ie. coating will maintain the performance with time). In most cases, an adaptation can be directly simulated by the estimation of correction (multiplicative) factors ($R$) that influence carbonation depth, diffusion coefficient, chloride concentration, critical chloride concentration or corrosion rate.

When the existing concrete cover is removed and replaced with new concrete, the deterioration process of the existing structure ceases and the deterioration process restarts at the time of cover replacement. It is assumed that (i) the new concrete cover re-passivates the steel reinforcement (corrosion rate is zero), (ii) corrosion loss of reinforcement is not rectified at time of cover replacement, and (iii) if corrosion initiates the corrosion rate is the same as for the original concrete cover (i.e. not affected by the new concrete cover).
3. CLIMATE CHANGE IMPACT ON CORROSION OF EXISTING CONCRETE BRIDGE STRUCTURES IN TEMPERATE CLIMATE ZONES

3.1 Chloride-Induced Corrosion of Concrete Bridges

Eleven existing bridges located within temperate climate zones in NSW are used as case studies of climate change impact on chloride-induced corrosion of the bridges. The bridges were chosen as representative of different construction periods at different regions, as shown in Table 3-1. The construction periods of bridges are divided into groups of prior to 1959, 1959-1970, 1971-1994, and post-1995. The data of post-1995 bridge structures is not available.

The profiles of chloride concentration were tested in 2008 at various locations of all listed bridges (RTA, 2008). The test data is then applied to calibrate the diffusion coefficient and surface chloride concentration, as indicated in the previous chapter, which are then used to project the corrosion until 2100 with the effect of climate change or not, as well as the influence of various adaptation measures.

Not all assessment will be reported here. More details can be found in the full report.
Table 3-1 List of bridges assessed for corrosion due to chloride penetration

<table>
<thead>
<tr>
<th>Construction Period</th>
<th>Southern</th>
<th>Sydney</th>
<th>Hunter</th>
<th>Northern</th>
</tr>
</thead>
<tbody>
<tr>
<td>Pre-1959</td>
<td><strong>Bridge BA1</strong></td>
<td><strong>Bridge BB1</strong></td>
<td><strong>Bridge BC1</strong></td>
<td><strong>Bridge BD1</strong></td>
</tr>
<tr>
<td></td>
<td>Built year: 1956</td>
<td>Built year: 1925</td>
<td>Built year: 1940</td>
<td>Built year: 1943</td>
</tr>
<tr>
<td></td>
<td>Exposure: C1</td>
<td>Exposure: C1</td>
<td>Exposure: C1 Sample: M1 (C2)</td>
<td>Exposure: C2</td>
</tr>
<tr>
<td></td>
<td>Cover: 55mm</td>
<td>Cover: 29mm</td>
<td>Cover: 50mm</td>
<td>Cover: 70mm</td>
</tr>
<tr>
<td>1959 – 1970</td>
<td><strong>Bridge BA2</strong></td>
<td><strong>Bridge BB2</strong></td>
<td><strong>Bridge BC2</strong></td>
<td><strong>Bridge BD2</strong></td>
</tr>
<tr>
<td></td>
<td>Exposure: C2</td>
<td>Exposure: C1</td>
<td>Exposure: C2</td>
<td>Exposure: C2</td>
</tr>
<tr>
<td></td>
<td>Cover: 50mm</td>
<td>Cover: 48mm</td>
<td>Cover: 53mm</td>
<td>Cover: 62mm</td>
</tr>
<tr>
<td>1971 – 1994</td>
<td><strong>Bridge BA3</strong></td>
<td>Test data not available</td>
<td><strong>Bridge BC3</strong></td>
<td><strong>Bridge BD3</strong></td>
</tr>
<tr>
<td></td>
<td>Exposure: C2</td>
<td>Exposure: C2</td>
<td>Exposure: C1/C2</td>
<td>Exposure: C1/C2</td>
</tr>
<tr>
<td></td>
<td>Cover: 58mm</td>
<td>Cover: 53mm</td>
<td>Cover: 67mm</td>
<td>Cover: 62mm</td>
</tr>
<tr>
<td>Post-1995</td>
<td>Test data not available</td>
<td>Test data not available</td>
<td>Test data not available</td>
<td>Test data not available</td>
</tr>
</tbody>
</table>

**Bridge BB1 (1925, Sydney)**

Concrete structures on bridges built in 1925, as shown in Figure 3-1. At a location with exposure C1, the probability of chloride-induced corrosion initiation and damage are estimated at 14% and 12% by 2008, respectively. The corresponding mean rebar loss is 0.25mm.

The probability of corrosion initiation is projected to be 70%, 68% and 67% for A1FI, A1B and 550 ppm stabilisation emission scenarios, respectively, in comparison with 62% in the absence of climate change, as shown in Figure 3-2. The probability of corrosion damage is projected to be 68%, 66% and 65% for the three emission scenario respectively in comparison with 60% in the absence of climate change. The mean rebar loss is 3.7mm, 3.5mm, and 3.4mm for the three emission scenarios respectively in comparison with 2.9mm in the absence of climate change. Therefore, the effect of climate change on the probability of chloride-induced corrosion initiation and damage is within 8% increase in probability value, (or 13% increase in percentage change) and the effect on the mean rebar loss is within 0.8mm increase by 2100, for the three emission scenarios.
Figure 3-1 Bridge BB1 constructed in 1925 and 1959, in Sydney (Source: RTA)

Figure 3-2 The probability of chloride-induced corrosion initiation and damage and mean rebar loss of Bridge BB1(1925) in a temperate climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘year 2000 level’ is the relevant value in the absence of climate change.
Bridges BC1 (1940, Hunter Region)

Bridge BC1 has been in service for 68 years to 2008, as shown in Figure 3-3, the tested structures are exposed to C1 and C2 class, respectively. The structure at a location of the bridge under exposure C1 shows very low corrosion in terms of the negligible probability of corrosion initiation and damage as well as mean rebar loss.

![Figure 3-3 Bridge BC1 constructed in 1940, in Hunter region of NSW (Source: RTA)](image)

The structure at a location with exposure C2, however, shows considerable likelihood of chloride-induced corrosion. The probabilities of corrosion initiation and damage were very high at 54% and 49%, respectively, and the mean rebar loss is 2.0mm by 2008. As shown in Figure 3-4, the probability of corrosion initiation and damage is above 99%, or almost certain by 2100. The corresponding mean rebar loss is 18mm in comparison with 16mm in the absence of climate change. The significance of the corrosion rate subdues the effect of climate change on corrosion initiation and damage that become almost certain anyway by 2100. It would be more important to consider the impact of climate change on maintenance measures that are applied to reduce the corrosion-induced structural deterioration. However, climate change is projected to cause up to 2.2mm increase in mean rebar loss by 2100, or 14% increase in comparison with the rebar loss in the absence of climate change.
Figure 3-4 The probability of chloride-induced corrosion initiation and damage and mean rebar loss of Bridge BC1 in a temperate climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘Year 2000 level’ is the relevant value in the absence of climate change.

**Bridge BD2 (1967, Northern Region)**

Bridge BD2 in NSW has been in service for 41 years by 2008, as shown in Figure 3-5. The tested part of structures at a location with exposure C2 shows very low corrosion by 2008. It is represented by the low probability of corrosion initiation and damage that was estimated around 3% while the mean rebar loss was negligible.

As shown in Figure 3-6, by 2100, the probability of chloride-induced corrosion initiation is projected to be 46%, 45% and 43% for A1FI, A1B and 550 ppm stabilisation emission scenario, respectively, in comparison with 39% in the absence of climate change. The probability of corrosion damage is projected to be 44%, 43% and 42% for the three emission scenario in comparison with 38% in the absence of climate change. The mean rebar loss is 4.6mm, 4.2mm and 4.0mm by 2100 for the three emission scenarios in comparison 3.3mm in the absence of climate change. In other words, climate change leads to 7% more in terms of change in probability value of corrosion initiation and damage (or 18% increase in percentage change of corrosion initiation probability due to climate change). It also causes up to 1.2mm increase in mean rebar loss (or 39% more than the mean rebar loss without the effect of climate change).
Figure 3-5 Bridge BD2 constructed in 1967 in the northern region of NSW (Source: RTA)

Figure 3-6 The probability of chloride-induced corrosion initiation and damage and mean rebar loss of Bridge BD2 in a temperate climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘year 2000 level’ is the relevant value in the absence of climate change.
3.2 Carbonation-Induced Corrosion of Concrete Bridges

Seven bridges exposed to C1 and C2 in the temperate climate zone in NSW, as shown in Table 3-2, are applied for the assessment of likely impact of climate change. Similar to those bridges for the case study of chloride-induced corrosion, the bridges are divided into four regions, i.e. Southern, Sydney, Hunter and Northern region, and four construction periods, i.e. pre-1959, 1959-1970, 1971-1994 and post-1995. The field carbonation test data was provided with a range of carbonation depth (CR) listed in the Table. However, they are not available for all bridges. The assessment will be done only for those with available carbonation depth information.

Not all assessment will be reported here. More details can be found in the full report.

Table 3-2. List of bridges assessed for carbonation-induced corrosion

<table>
<thead>
<tr>
<th>Construction Period</th>
<th>Southern</th>
<th>Region</th>
<th>Sydney</th>
<th>Hunter</th>
<th>Northern</th>
</tr>
</thead>
<tbody>
<tr>
<td>Pre-1959</td>
<td>Bridge BA1 1956</td>
<td>Bridge BB1 1925</td>
<td>Test data not available</td>
<td>Test data not available</td>
<td></td>
</tr>
<tr>
<td>CR: 10 – 25 mm</td>
<td>CR: 5 – 35 mm</td>
<td>Cover: 55mm</td>
<td>Cover: 29mm</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Cover: 50mm</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>1959 – 1970         Bridge BA2 1959</td>
<td>Bridge BB2 1959</td>
<td>Bridge BC2 1968</td>
<td>Test data not available</td>
<td></td>
<td></td>
</tr>
<tr>
<td>CR: 1 – 5 mm</td>
<td>CR: 5 – 15 mm</td>
<td>CR: 5 – 25 mm</td>
<td>Test data not available</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Cover: 50mm</td>
<td>Cover: 48mm</td>
<td>Cover: 53mm</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>1971 – 1994         Bridge BA3 1980</td>
<td>Test data not available</td>
<td>Test data not available</td>
<td>Bridge BD3 1984</td>
<td></td>
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<tr>
<td>CR: 5 – 20 mm</td>
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<tr>
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<td></td>
<td>Cover: 62mm</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Post-1995           Test data not available</td>
<td>Test data not available</td>
<td>Test data not available</td>
<td>Test data not available</td>
<td></td>
<td></td>
</tr>
</tbody>
</table>

Bridge BA1 (1956, Southern Region)

Bridge BA1 was built in 1956. Concrete in the test site has cover of 55mm with exposure C1-C2. The mean carbonation depth in 2008 is 17.5mm as tested. As shown
in Figure 3-9, climate change may have noticeable impact on the carbonation depth. By 2100, the mean carbonation depth is 34.0mm for A1FI emission scenario, 31.9mm for A1B emission scenario, and 30.1mm for 550ppm stabilisation scenario in comparison with 26.5mm estimated in the absence of climate change, or 28%, 20% and 14% more, respectively. However, 55mm cover of concrete has almost prevented the occurrence of corrosion initiation and damage represented by their negligible probability along with a very small mean rebar loss, even if climate change would contribute a significant increase in carbonation depth. On the other hand, it demonstrated the importance of proper cover design to reduce the impact of carbonation on corrosion of reinforcement.

Bridge BA1 (1925, Sydney)

The part of structures of Bridge BA1 built in 1925 has a concrete cover of 29mm. Field tests found carbonation in the range of 5-35mm, very likely inducing corrosion. As shown in Figure 3-10, the mean carbonation depth is 35.2mm for A1FI emission scenario, 33.3mm for A1B emission scenario, and 31.6 for 550ppm stabilisation emission scenario by 2100, in comparison with 28.4mm estimated without the effect of climate change.
As carbonation at the mean depth of 28.4mm in the absence of climate change is already very close to cover thickness of 29mm, any increase in carbonation depth caused by climate change may significantly impact on the probability of corrosion initiation and damage. As shown in Figure 3-10, the probability of corrosion initiation is 72%, 67% and 63% for the three emission scenarios respectively in comparison with 51% estimated in the absence of climate change. In other words, the probability increases 21%, 16% and 12% in value, or an increase of 41%, 31% and 14% in percentage change due to climate change. Meanwhile, the probability of corrosion damage is 63%, 60% and 55% in comparison with 44% estimated in the absence of climate change.
4. **CLIMATE CHANGE IMPACT ON CORROSION OF EXISTING CONCRETE PORT STRUCTURES IN TROPICAL CLIMATE ZONES**

Concrete structures, including a slab soffit and a column in a berth of Port Townsville, Queensland, as shown in Figure 4-1, are considered for the case studies.

The berth was constructed circa 1960. It consists of a flat reinforced concrete deck approximately 530 mm thick, supported by circular reinforced concrete columns, as shown in Figure 4-2 (or the head photo of this chapter), generally 760 mm in diameter, with some 910 mm in diameter, on a 6.1 m square grid which are founded on pile caps which are in turn supported by driven universal piles. The berth and its adjacent have been constructed as a single rectangular wharf structure with the central portion of the concrete structure founded of fill. They appear to comprise the same reinforced concrete retaining walls, piles and slabs that extend around three sides of the wharf. The total wharf structure is approximately 250 m long and 70 m wide with the piled area for each berth is approximately 22m wide.

Following detailed inspections in 2008, the reinforced concrete of the berth is in a relatively good condition considering its age of 50 years (48 years in 2008). The results of the 2008 condition assessment indicated that with ongoing inspection and maintenance the deck should be serviceable for the next 50 years. The support columns for the berth were at risk of widespread deterioration within the next 10 years and are currently undergoing remedial works.
The test was done in 37 locations of the slab soffit with a cover of 50mm, and 8 locations of columns with a cover of 60mm. The test data of chloride concentration profiles and carbonation depth were provided by Port of Townsville, (2009). While chloride-induced corrosion is assessed on the location of slab soffits or columns with the highest chloride concentration, the carbonation-induced corrosion is assessed on the whole structures of the berth with different carbonation depths considered as an uncertainty in carbonation.
4.1 Chloride-Induced Corrosion of Port Structures

**Slab Soffits**

The tests were done for the concrete slab soffit between column 33 and 34 at column line N of the berth in Port Townsville in 2008. The slab has a cover of 50mm and is considered under exposure C2. The chloride content analysis shows the chloride concentration at 0.32%, 0.21%, 0.12% and 0.04 % by weight of concrete in the depth of 5mm, 20mm, 40mm and 50mm, respectively. With the concentration profile, the probability of chloride-induced corrosion initiation and damage is estimated at 10% and 9%, respectively, and likely mean rebar loss at 0.34mm.

Figure 4-3 Probability of chloride-induced corrosion initiation and damage and mean rebar loss of concrete structures of a slab soffit (33-34N) in a tropical climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘Year 2000 level’ is the relevant value in the absence of climate change.

As shown in Figure 4-3, the probability of corrosion initiation is projected to reach 59% by 2100 for A1FI emission scenario, 58% for A1B emission scenario and 57% for 550ppm stabilisation emission scenario in comparison with 55% in the absence of climate change. The probability of corrosion damage is 58%, 57% and 56% for the three emission scenarios respectively in comparison with 54% estimated without the effect of climate change. The mean rebar loss is project to be 9.3mm, 9mm and 8.8mm.
for three emission scenarios in comparison with 8mm without the effect of climate change.

In this regard, the impact of climate change by 2100 is limited within 4% increase in probability value of corrosion initiation and damage (or an increase of 7.4% in percentage), and it is within 1.3mm of mean rebar loss or a increase of 16%.

Columns

Column L5 of the berth in Port Townsville is selected as the case study of concrete port structures. It has cover of 60mm. The core sample from the column for chloride concentration analysis is located 500mm below tapered section of the column, and is considered at exposure C2. The chloride concentration analysis was done in 2008 and details the concentration profile that is 0.28%, 0.1%, 0.04%, 0% by the weight of concrete at the depth of 5mm, 20mm, 40mm and 60mm, respectively. Based on the chloride concentration profile, the probability of corrosion initiation and damage is less than 1% by 2008 and the mean rebar loss is estimated at 0.024mm.

As shown in Figure 4-4, the probability of corrosion of initiation and damage is projected less than 1.2% for all emission scenarios. The significance of climate change impact is very limited.

4.2 Carbonation-Induced Corrosion of Port Structures

Slab Soffits

The tests were done at 37 locations for the concrete slab soffit of the berth in Port Townsville in 2008. The slab has a cover of 50mm. The test of core sample from the slab shows carbonation depth in the range of 22mm to 55mm in 2008. On the basis of the test, it is very unlikely that there would be any occurrence of carbonation-induced corrosion initiation and damage by 2008.

As shown in Figure 4-5, the probability of corrosion initiation and damage as well as mean rebar loss may increase significantly over the years, and climate change may even magnify the change.
As seen in the figure, by 2100, the probability of carbonation-induced corrosion initiation is projected to reach 91% for A1FI emission scenario, 89% for A1B emission scenario, and 86% for 550 ppm stabilisation emission scenario in comparison with 76% in the absence of climate change. In other words, climate change causes 15%, 13% and 10% increase in probability value. In the term of percentage change, the climate change leads to an increase of 20%, 17% and 13%, respectively. In fact, climate change leads to the largest impact on the probability of corrosion initiation around 2080, where the probability is 16% more than one estimated without the effect of climate change, or an increase of 31% in the term of percentage change.
Figure 4-5 Carbonation depth, probability of carbonation-induced corrosion initiation and damage and mean rebar loss of concrete structures of a column (L5) in a tropical climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘Year 2000 level’ is the relevant value in the absence of climate change.

Meanwhile, the probability of corrosion damage is 78%, 76% and 73% by 2100 for the three emission scenario, respectively, in comparison with 62% in the absence of climate change. In the term of percentage change, it is an increase of 26%, 23% and 18%, respectively. The mean rebar loss is 1.1mm, 1.0mm, 0.97mm for the three emission scenario in comparison with 0.79mm estimated without the effect of climate change.

Columns

Columns of the berth have a cover of 60mm. The core samples of the column were taken from 7 locations for carbonation depth assessment. The samples are located 500mm below tapered section of the column, showing a carbonation depth in the range of 22mm to 38mm by 2008. At such a depth of carbonation, it is very unlikely that the carbonation would cause any corrosion initiation and damage. As shown in Figure 4-6, without the effect of climate change, the carbonation depth is projected to reach 46.1mm by 2100. Although it is approaching the end of cover (60mm), it would still not cause significant change in the probability of corrosion initiation and damage.
As shown in the figure, the consideration of climate change significantly increases the probability of carbonation-induced corrosion, especially after 2060 for corrosion initiation and 2070 for corrosion damage. The probability of corrosion initiation is 45% for A1FI emission scenario, 28% for A1B emission scenario and 14% for 550 ppm stabilisation emission scenario, in comparison with only 1% in the absence of climate change. The probability of corrosion damage is projected to be 16%, 8.8% and 4.0% for the three emission scenarios in comparison with 0.17% estimated without the effect of climate change.

As a result, in this case, the climate change may have considerable impact on the reliability of concrete port structures. It is noted that the design of concrete cover is very critical to prevent from or reduce the impact of climate change.

Figure 4-6 Carbonation depth, probability of carbonation-induced corrosion initiation and damage and mean rebar loss of concrete structures of a column (L5) in a tropical climate zone, with the effect of climate change in comparison with the baseline in the absence of climate change. ‘Year 2000 level’ is the relevant value in the absence of climate change.
5. ADAPTATION MEASURES TO COUNTERACT CLIMATE CHANGE IMPACT

Climate adaptation is considered as a process of deliberate change in anticipation of or in reaction to climatic stimuli and stress, in order to moderate or reduce adverse effects onto the system of our interest (IPCC, 2007). The critical component in climate adaptation is the adaptive capacity that measures the capability of the system to maintain its functionality and integrity under the disturbance of external stimuli and stresses. Enhancement of climate adaptive capacity, along with reducing vulnerability, is considered as one of prime approaches to counteract the changing climate.

For existing concrete structures under changing climate, the climate adaptation in terms of the enhancement of adaptive capacity can be done by developing new technologies for maintenance to counter the impact of increasing corrosion risk under changing climate. On the other hand, there is a wide range of conventional maintenance and retrofitting options that can enhance the durability of concrete structures and these can be applied to reduce the adverse affects of climate change. The maintenance options generally include surface coating, realkalisation, chloride extraction, cathodic protection, and cover replacement.

Creating a surface barrier by coating is more appropriate for reducing the exposure of concrete structure to external stimuli. Meanwhile, extraction and cathodic protection is more commonly carried out for structures with high corrosion risk to reduce the ingress of deleterious agents. The cover replacement is most effective, but also the most expensive options, followed by cathodic protection that also has operating cost and then
realkanisation or chloride extraction. The surface coating is the cheapest option, but is also less effective.

5.1 Simulation of Adaptation Measures for Chloride- and Carbonation-Induced Corrosion

To consider the worst scenario of climate change impact, A1FI emission scenario is applied to assess the adaptation of concrete structures for chloride-induced corrosion. Meanwhile, as indicated previously, the simulation of adaptation measures for chloride- and carbonation-induced corrosion is represented by various factors.

It is assumed in the simulation of adaptation measures for chloride-induced corrosion:

1. Electrochemical Chloride Extraction 90% at a cost of about $600/m². It is a permanent solution if a surface coating is applied and properly maintained.
2. Polyurethane sealer at a cost of about $40/m². The coating is applied every 15 years.
3. Polymer-modified (p-m) cementious coating at a cost is about $40/m². The coating is applied every 15 years.
4. Replacing existing cover at a cost of about $2,500/m².
5. Cathodic protection at a cost including $800/m² initial cost and $10/m²/year operating cost. It permanently stops corrosion initiation and damage.

At the same time, it is assumed in the simulation of adaptation measure for carbonation-induced corrosion:

6. Realkalisation at a cost of about $600/m². It is a permanent solution if a surface coating is applied and properly maintained.
7. Acrylic-based coating: at a cost of about $50/m². The coating is applied every 15 years.
8. Replacing existing cover at a cost of about $2,500/m².

Adaptation measures are assumed to be applied in 2011 in the simulation, and all cost is the present value in 2011.

5.2 Cost/Benefit Assessment of Adaptation Options

Since there is lack of information of damage cost, a proxy of the benefit due to the implementation of adaptation options in changing climate is defined by:

\[
\text{Benefit} = \text{Risk(BAU | climate change)} - \text{Risk(adaptation | climate change)} \quad (5-1)
\]

where Risk(BAU | climate change) is the risk in the circumstance of ‘business as usual’ or ‘do-nothing’ under climate change, and Risk(adaptation | climate change) is the residual risk after the implementation of an adaptation option under climate change.
In fact, the benefit consists of three parts, 1) mitigate corrosion risk of corrosion completely due to climate change, 2) increase corrosion resistance or adaptive capacity by the maintenance in the absence of climate change (the maintenance option is the same as adaptation option, applied to mitigate corrosion risk as usual but nothing to do with climate change), and 3) offset loss of the benefit due to the reduced effectiveness of the adaptation option as a result of climate change. Meanwhile, the adaptation cost is estimated on the basis of initial and operating cost to maintain the performance or function of adaptation measures given a discount rate.

In addition, ‘Benefit’ as described in Eqn.(5-1) also represents the degree of the reduction of corrosion risk when an adaptation option is applied, and can thus be considered as the representation of adaptation effectiveness. An option is not recommend when ‘Benefit≤0’. The maximum benefit or adaptation effectiveness can be achieved when the risk, or Risk(adaptation | climate change) in Eqn.(5-1), approaches zero after an adaptation option implemented.

Meanwhile, based on the cost and the effectiveness, an adaptation effectiveness diagram, such as shown in Figure 5-1, can be established to inform decision-makers to select proper adaptation options based on both effectiveness and cost. In the diagram, each bar represents an adaptation option with its width representing adaptation effectiveness and height representing cost. The optimal option is low cost and great effectiveness.

Figure 5-1 Conceptual illustration of adaptation effectiveness diagram
### 5.3 Adaptation Assessment for Chloride-Induced Corrosion and Cost/Benefit

**Bridge BB1 Constructed in 1925 in Sydney**

Figure 5-2 describes the effect of adaptation measures (1)-(4) described above on the probability of corrosion initiation and damage as well as rebar loss. It appears the most effective approach is the replacement of old 29mm cover of concrete with new concrete that meets the current design stands of 65mm cover at exposure C2. The measure not only counteract the impact of climate change, but also enough to maintain the probability of corrosion initiation and damage at a very low level and almost no further rebar loss.

![Figure 5-2](image)

(a) Corrosion initiation probability  
(b) Corrosion damage probability  
(c) Mean rebar loss (mm)

Figure 5-2 Probability of chloride-induced corrosion initiation and damage and mean rebar loss of concrete structures of Bridge BB1 (1925) in NSW considering adaptation options (1)-(4). ‘Impact’ - with the effect of climate change; ‘Year 2000 level’ – the relevant value in the absence of climate change.

The second most effective approach is chloride extraction that assumes that 90% of chloride was removed. Although it is still projected that there is an increase in the probability of corrosion initiation and damage, but it maintain the probability of corrosion initiation below the estimation in the absence of climate change until 2080, and maintain the probability of corrosion damage below the projection in the absence of climate change. In another word, it has at least mitigated the impact of climate change.
The use of p-m cementious coating may also maintain the probability of corrosion initiation and damage below the estimation in the absence of climate change, while relatively less effective. In addition, polyurethane sealer, as indicated in Table 2-2, has little effect on the probability and mean rebar loss, and is thus not recommended for use.

Considering a service life of concrete structure till 2100, the cost and cost/benefit ratio is shown in Figure 5-3. It should be pointed out that the cost/benefit ratio is not in the sense of the ratio of dollar to dollar, and it is just a relative indicator for the comparison. As shown in the figure, cover replacement is most expansive and also very high in cost/benefit ratio among the five options, while p-m cementious coating is the least. The cost/benefit ratio of the implementation of polyurethane sealer is most sensitive to the discount rate.

![Figure 5-3](image)

Figure 5-3 Cost and cost/benefit in relation to discount rate for the implementation of adaptation options for chloride-induced corrosion of concrete structures of Endeavour Bridge (1925) in NSW under climate change.

The cost is contributed to reduce corrosion damage risks as usual and the increased risks due to climate change as well. On the basis of Eqn.(5-3), the cost contributes to three part, 1) mitigating corrosion risk due to climate change, 2) increasing corrosion resistance, and 3) offsetting the loss of adaptation effectiveness due to climate change. Table 5-1 gives the subdivision of the cost related to the three parts.

For an example, for p-m cementious coating, 31% of the total cost is contributed to the mitigation of increased corrosion risk due to climate change, 46% is contributed to improve the resistance of concrete structures to chloride attack, and 23% is contributed to the loss of the effectiveness of adaptation options due to climate change. For polyurethane, nearly a half of cost is actually for offsetting the loss of its effectiveness due to climate change.

It should be pointed out that the cathodic protection in the table is assumedly not affected by the climate change. In fact, its effectiveness may be affected by sea level rise, which will not be discussed in this report.
Finally, Figure 5-4 indicates the overall effectiveness of adaptation options under changing climate in relation to the cost considering a discount rate of 0.03. The width of the bar represents the relative adaptation effectiveness of each option. Polyurethane sealer is least effective.

Table 5-1 Subdivision of the cost of the adaptation/maintenance measures for chloride-induced corrosion of Bridge BB1 (1925) in Sydney

<table>
<thead>
<tr>
<th>Adaptation/Maintenance Measures</th>
<th>Percentage of total cost to mitigate increased risk due to climate change</th>
<th>Percentage of the total cost to increase corrosion resistance</th>
<th>Percentage of the total cost to offset the loss of adaptation effectiveness due to climate change</th>
</tr>
</thead>
<tbody>
<tr>
<td>Polyurethane sealer</td>
<td>47%</td>
<td>7%</td>
<td>46%</td>
</tr>
<tr>
<td>p-m cementious coating</td>
<td>31%</td>
<td>46%</td>
<td>23%</td>
</tr>
<tr>
<td>Chloride extraction</td>
<td>29%</td>
<td>53%</td>
<td>18%</td>
</tr>
<tr>
<td>Replacement</td>
<td>22%</td>
<td>78%</td>
<td>0%</td>
</tr>
<tr>
<td>Cathodic protection</td>
<td>22%</td>
<td>78%</td>
<td>0%</td>
</tr>
</tbody>
</table>

Figure 5-4 Cost and effectiveness of adaptation options for chloride-induced corrosion of concrete structures of Bridge BB1 (1925) in NSW, at a discount rate of 0.03.
Bridges BD2 Constructed in 1967 in Northern Region

Figure 5-5 describes the effect of adaptation measures (1)-(4) described above on the probability of corrosion initiation and damage as well as rebar loss. It appears the most effective approach is the use of p-m cementious coatings on the surface of 62mm cover of concrete that already meets the current design stands for structures at exposure C2. The measure not only counteracts the impact of climate change, but also enough to maintain the probability of corrosion initiation and damage at a reasonably low level.

The second most effective approach is cover replacement that assumes the same cover thickness but at lower rate of penetration. Although it is still projected that there is an increase in the probability of corrosion initiation and damage, but it maintains the probability of corrosion initiation below the estimation in the absence of climate change until 2080, and maintains the probability of corrosion damage below the projection in the absence of climate change. In others word, it has at least mitigated the impact of climate change.
Considering a service life of concrete structure till 2100, the cost and cost/benefit ratio is shown in Figure 5-3. Once again, the ratio is just a relative indicator for the purpose of comparison. As shown in the figure, cover replacement is most expensive and also highest cost/benefit among the five options, while p-m cementious coating is the least. Different from the last case, the use of polyurethane coating appears also much lower in addition to its lower sensitivity to the discount rate.

Table 5-2 gives the subdivision of the cost related to the three parts. Comparing to the previous case, polyurethane sealer is much more effective with 29% of the total cost is contributed to the mitigation of increased corrosion risk due to climate change, 57% is contributed to improve the resistance of concrete structures to chloride attack, and moderate 14% is contributed to the loss of the effectiveness of adaptation options due to climate change. In general, more than a half of the total cost goes to increase the resistance to corrosion.
Once again, it should be pointed out that the cathodic protection in the table is assumedly not affected by the climate change. However, its effectiveness may be affected by sea level rise, which will not be discussed in this report.

In addition, Figure 5-7 indicates the overall effectiveness of adaptation options under changing climate in relation to the cost with considering a discount rate of 0.03. In the figure, the width of each bar represents the relative adaptation effectiveness of each option, while the height represents the cost. For this case, p-m cementious coating becomes a very attractive option with low cost and high effectiveness. On the other hand, cover replacement is least attractive with high cost and relatively low effectiveness.

![Figure 5-7 Cost and effectiveness of adaptation options for chloride-induced corrosion of concrete structures of Bridge BD2(1967) in the northern region of NSW, at a discount rate of 0.03.](image)

**Concrete Slab Soffits in Port Townsville**

Figure 5-8 describes the effect of adaptation measures (1)-(4) described above on the probability of corrosion initiation and damage as well as rebar loss of concrete slabs in one of berths in Port of Townsville. It appears the most effective approach to reduce the corrosion is the use of either cover replacement or p-m cementious coatings, though the probability corrosion initiation and damage increase again to almost 30% by 2100 under A1FI emission scenario.
Other approaches including chloride extraction that only marginally reduces the probability of corrosion initiation and damage. The use of polyurethane sealer cannot even maintain the probability below the estimation in the absence of climate change.

Figure 5-8 Probability of chloride-induced corrosion initiation and damage and mean rebar loss of structures of concrete slab in Port of Townsville considering adaptation options (1)-(4). ‘Impact’ - with the effect of climate change; ‘Year 2000 level’ – the relevant value in the absence of climate change.

Considering a service life of concrete slab structure till 2100, the cost and cost/benefit ratio is shown in Figure 5-9. As seen, cover replacement is most expansive and also highest cost/benefit among the five options, while p-m cementious coating is the least. It also indicates that the cost/benefit ratio to use polyurethane coating appears sensitive to the discount rate.

Table 5-3 gives the subdivision of the cost related to the three parts. As seen, polyurethane sealer show less effective with 36% of the total cost contributed to the loss of the effectiveness of adaptation options due to climate change. Meanwhile, 38% of the total cost is contributed to the mitigation of increased corrosion risk due to climate change, and 25% is contributed to improve the resistance of concrete structures to chloride attack. In general, except of polyurethane sealer, more than a half of the total cost of all options goes to increase the resistance to corrosion, especially cathodic protection and p-m cementious coating that is as high as 87% and 75%, respectively. In
another word, the adaptive capacity will be significantly strengthened to resist chloride-induced corrosion.

Figure 5-9 Cost and cost/benefit in relation to discount rate for the implementation of adaptation options for chloride-induced corrosion of a concrete slab in Port of Townsville under climate change.

Table 5-3 Subdivision of the cost of the adaptation/maintenance measures for chloride-induced corrosion of concrete slab of port structure in the port of Townsville.

<table>
<thead>
<tr>
<th>Adaptation/Maintenance Measures</th>
<th>Percentage of total cost to mitigate increased risk due to climate change</th>
<th>Percentage of the total cost to increase corrosion resistance</th>
<th>Percentage of the total cost to offset the loss of adaptation effectiveness due to climate change</th>
</tr>
</thead>
<tbody>
<tr>
<td>Polyurethane sealer</td>
<td>38%</td>
<td>25%</td>
<td>36%</td>
</tr>
<tr>
<td>p-m cementious coating</td>
<td>16%</td>
<td>75%</td>
<td>9%</td>
</tr>
<tr>
<td>Chloride extraction</td>
<td>28%</td>
<td>51%</td>
<td>21%</td>
</tr>
<tr>
<td>Replacement</td>
<td>20%</td>
<td>71%</td>
<td>8%</td>
</tr>
<tr>
<td>Cathodic protection</td>
<td>13%</td>
<td>87%</td>
<td>0%</td>
</tr>
</tbody>
</table>

As discussed before, the cathodic protection in the table is assumedly not affected by the climate change, which may indeed be affected by sea level rise. However, it will not be discussed in this report.

Figure 5-10 describes the overall effectiveness of adaptation options under changing climate in relation to the cost with considering a discount rate of 0.03. In the figure, the width of each bar represents the relative effectiveness of each adaptation option, while the height represents the cost. For this case, cathodic protection and p-m cementious coating becomes two options to effectively mitigate chloride-induced corrosion under changing climate.
5.4 Adaptation Assessment for Carbonation-Induced Corrosion and Cost/Benefit

Figure 5-11 describes the effect of adaptation measures (6)-(8) introduced above on the probability of corrosion damage of concrete slabs in one of berths in Port of Townsville. Considering that the acrylic-based surface coating is normally applied in dry environment, only two options, realkanisation and cover replacement, are considered for the concrete slab at exposure C2. Realkalisisation is an electrochemical process to raise pH near reinforcement. After realkalisisation, concrete will not easily recarbonate. It is therefore assumed that corrosion will stop permanently. Meanwhile, cover replacement also stops corrosion while it may not stop carbonation.

Considering a service life of concrete slab structure till 2100, the cost and cost/benefit ratio is shown in Figure 5-9. As seen, cover replacement is most expansive and also highest cost/benefit among the five options, while p-m cementious coating is the least. It also indicates that the cost/benefit ratio to use polyurethane coating appears sensitive to the discount rate.
Figure 5-11 Probability of carbonation-induced corrosion damage of structures of concrete slab in Port of Townsville considering adaptation options (6) and (8). ‘Impact’ - with the effect of climate change; ‘Year 2000 level’ – the relevant value in the absence of climate change.

Table 5-4 gives the subdivision of the cost related to the three parts. As seen, both alkanisation and cover replacement 38% of the total cost contributed to the mitigation of increased corrosion risk due to climate change, and other 62% is contributed to improve the resistance of concrete structures to chloride attack or adaptive capacity. However, their cost to reach that effectiveness is different.

Figure 5-13 describes the overall effectiveness of adaptation options under changing climate in relation to the cost with considering a discount rate of 0.03. In the figure, the width of each bar represents the relative effectiveness of each adaptation option for
carbonation-induced corrosion, while the height represents the cost. For this case, realkanisation has the advantage over the cover replacement to effectively mitigate chloride-induced corrosion under changing climate due to its low cost.

Table 5-4 Subdivision of the cost of the adaptation/maintenance measures for carbonation-induced corrosion of concrete slab of port structure in the port of Townsville.

<table>
<thead>
<tr>
<th>Adaptation/Maintenance Measures</th>
<th>Percentage of total cost to mitigate increased risk due to climate change</th>
<th>Percentage of the total cost to increase corrosion resistance</th>
<th>Percentage of the total cost to offset the loss of adaptation effectiveness due to climate change</th>
</tr>
</thead>
<tbody>
<tr>
<td>Realkanisation</td>
<td>38%</td>
<td>62%</td>
<td>0%</td>
</tr>
<tr>
<td>Cover replacement</td>
<td>38%</td>
<td>62%</td>
<td>0%</td>
</tr>
</tbody>
</table>

Figure 5-13 Cost and effectiveness of adaptation options for carbonation-induced corrosion of concrete slab structures in the Port of Townsville at a discount rate of 0.03.
6. SUMMARY

Since most designs of existing concrete structures do not take into account the effect of a changing environment, there are concerns that many existing concrete infrastructure are likely to suffer from decreased durability, safety and serviceability.

This report has established a probabilistic simulation approach based on chloride- and carbonation-induced corrosion models that can be calibrated by the historical field measurement of chloride concentration and carbonation penetration and then can project the probability of corrosion initiation and damage into future. We developed models of adaptation options and looked at their effects on various parameters of corrosion models for existing concrete structures, such as carbonation depth and carbon dioxide diffusion coefficient for carbonation-induced corrosion, and chloride concentration, chloride diffusion coefficient, critical chloride concentration for chloride-induced corrosion, and corrosion rate. These models were then applied to simulate the corrosion under the influence of adaptation measure, including surface coating, alkanisation, chloride extraction, cover replacement, cathodic protection and so on. The effect of climate change is also considered in the models through the influence of carbon dioxide concentration, temperature and relative humidity.

Two typical concrete structures, bridges and port structures, are used as case studies to elucidate the impact of climate change on existing concrete structures and available options as well as their effectiveness in reducing the impact of climate change and increasing adaptive capacity to counteract corrosion. The investigated bridges are located in Sydney, Southern, Northern and Hunter regions of NSW and constructed during the period of pre-1959, 1959-1970, 1971-1994 and post-1995. At the same time, the investigated port structures are concrete slab soffits and columns from a berth that is managed by Port of Townsville Limited. As found, although all investigated structures
are under exposure C2, the concrete cover does not necessarily meet the current standard, such as AS 3600-2009, that requires 65mm. In fact, the cover of the most of existing bridges in this case study is less than 65mm, with one reaching as low as 29mm. This explained more on the importance to specifically assess durability of existing bridges and especially when considering the impact of climate change on the durability in relation to carbonation- and chloride-induced corrosion.

In the part 2 of the report, it is indicated that the effect of climate change on chloride-induced corrosion of concrete structures which design follows the AS3600 and AS5100.5, is within an increase of 3.5% in probability value by 2100. In practice, the change of the probability can go higher due to non-binding on the standards or lack of quality assurance in construction, for example, use of lower concrete cover. The bridges in Sydney, constructed in 1925 with 29mm cover, shows that the climate change at A1FI emission scenario can leads up to 8% increase in corrosion initiation and damage probability value, or 13% increase in percentage change. Even for a modern bridge in the northern region of NSW, constructed in 1984, the structure on the bridge at exposure C1 and C2 may experience up to 5-7% increase in probability value by 2100 in the presence of climate change impact at A1FI emission scenario.

Due to an improper concrete cover of some of the early constructed bridges, climate change may lead to a considerable impact on carbonation-induced corrosion. For example, the bridge constructed in 1925 in Sydney has only 29mm concrete cover. The probability of corrosion initiation is 72%, 67% and 63% by 2100 for A1FI, A1B and 550 ppm stabilisation emission scenarios, respectively, in comparison with 51% estimated in the absence of climate change. In another word, the probability increases 21%, 16% and 12% in value, or an increase of 41%, 31% and 14% in percentage change due to climate change. Meanwhile, the probability of corrosion damage is 63%, 60% and 55% in comparison with 44% estimated in the absence of climate change, also a significant increase due to climate change. For a concrete column in Port Townsville, climate change impact may lead to an increase of corrosion initiation probability form 1% to 45% for A1FI emission scenario, 28% for A1B emission scenario and 14% for 550 ppm stabilisation emission scenario by 2100. It also leads to the increase corrosion damage probability from 0.2% to 16%, 8.8% and 4.0% for the three emission scenarios by 2100.

Both chloride-induced and carbonation-induced corrosion show the potential experience of a scalable impact of climate change, which should be considered for maintenance planning. Adaptation options should also be developed and optimised to mitigate the impact and enhance the adaptive capacity of concrete structures to changing climate.

In the simulation of implementation of adaptations to counteract the impact of climate change, five options including electrochemical chloride extraction, polyurethane sealer, polymer-modified cementious coating, cover replacement and cathodic protection, were considered to reduce chloride-induced corrosion. Other two options including realalkalisation and cover replacement were introduced to mitigate carbonation-induced corrosion. Meanwhile, cost and adaptation effectiveness are also introduced to quantify the adaptation options for the identification of the most preferable option in association with a specific concrete structure, such as a slab or column. The cost includes initial
implementation cost and on-running operating cost, which are all converted to their present value in 2011 with a discount rate ranging from 0.01 to 0.1 selected for sensitivity assessment. The effectiveness, also known as a proxy of benefit due to the implementation of adaptation options, is defined as the amount of reduction in corrosion risk from ‘business as usual (BAU)’ after implementing adaptations. As a result, adaptation effectiveness diagram is developed by combining cost with adaptation effectiveness, which may facilitate the decision-making in developing adaptation strategies that maximise adaptation effectiveness with minimised cost.

The cost contributes to three factors, i.e. 1) reducing the impact of climate change, 2) increasing adaptive capacity to resist corrosion, and 3) offsetting the loss of adaptation effectiveness due to climate change. In general, the more there is more on offsetting, less effective is the option. The case study of concrete bridges indicates that the replacement of concrete cover is often to be the most effective options, but it is also the most expensive one. Surface coating is the least costly, but is usually, but not always relatively less effective. For the bridge constructed in 1925 in Sydney, cathodic protection is the preferred adaptation measure to mitigate the chloride-induced corrosion damage due to its greater effectiveness and moderate cost. Among the total cost, 22% is contributed to mitigate the increase corrosion damage risk due to climate change, 78% is contributed to increase adaptive capacity to resist corrosion, and nothing is contributed to offset the loss of adaptation effectiveness due to climate change. It should be pointed out the effectiveness of cathodic protection is in fact affected by sea level rise that may change the cost for offset the loss of effectiveness. At the same time, the use of polyurethane sealer is the least preferred due to its very low effectiveness though low cost. For this, 46% of the total cost is contributed to offset the loss, which is not really beneficial to the enhancement of adaptive capacity to counteract corrosion damage.

Depending on residual risk of corrosion damage of concrete structure after implementing adaptation options, the preferred adaptation option can vary. For a bridge constructed in 1967 in the northern region of NSW, polymer-modified cementious coating appears the most preferable due to its great effectiveness with 24% of the total cost is contributed to mitigate the increasing risk as a result of climate change, 57% is contributed to strengthen the adaptive capacity, and only 4% is contributed to offset the loss of adaptation effectiveness, in comparison with 14% for polyurethane sealer, 10% for cathodic protection. Meanwhile, it has a cost much lower than cover replacement, cathodic protection and chloride extraction.

A similar approach was applied for the cost/benefit assessment of adaptation options for carbonation-induced corrosion of port structures in relation to realkalisation and cover replacement are considered.

As reported in part 2 of the report, the climate change impact assessment on the aspect of design that follows the Australian standards may provide general rules for concrete structural design taking into account effects of changing climate. Different from that, impacts on existing concrete infrastructure and the adaptation that should be applied to mitigate the impact are specific due to the uniqueness of individual structures especially regarding their different local environment exposure history as well as the uncertainties
in construction and maintenance. Therefore, an effective adaptation option should be developed at the level of individual concrete structures. Finally, cost and benefit assessment should further be developed to consider the lifecycle of concrete infrastructure.

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